

広島大学学術情報リポジトリ
Hiroshima University Institutional Repository

Title	High and stable substrate specificities of microorganisms in enhanced biological phosphorus removal plants
Author(s)	Kindaichi, Tomonori; Nierychlo, Marta; Kragelund, Caroline; Nielsen, Jeppe Lund; Nielsen, Per Halkjær
Citation	Environmental Microbiology , 15 (6) : 1821 - 1831
Issue Date	2013-06
DOI	10.1111/1462-2920.12074
Self DOI	
URL	https://ir.lib.hiroshima-u.ac.jp/00041294
Right	Copyright (c) 2012 John Wiley & Sons Ltd and Society for Applied Microbiology
Relation	



1 **Revised**

2 **MS for Environmental Microbiology**

3 **Article Type:** Research Article

4

5 **High and stable substrate specificities of microorganisms in enhanced biological**
6 **phosphorus removal plants**

7

8

9 **Tomonori Kindaichi^{a,b}, Marta Nierychlo^b, Caroline Kragelund^b, Jeppe Lund Nielsen^b,**

10 **Per Halkjær Nielsen^{b,*}**

11

12 ^a Department of Civil and Environmental Engineering, Graduate School of Engineering, Hiroshima
13 University, 1-4-1 Kagamiyama, Higashihiroshima, 739-8527 Japan.

14 ^b Department of Biotechnology, Chemistry and Environmental Engineering, Aalborg University,
15 Sohngaardsholmsvej 49, DK-9000 Aalborg, Denmark.

16

17 Running Title: Substrate Specificities of Microorganisms in EBPR Plants

18

19 * Corresponding author:

20 Per Halkjær Nielsen

21 Department of Biotechnology, Chemistry and Environmental Engineering, Aalborg University,
22 Sohngaardsholmsvej 49, DK-9000 Aalborg, Denmark.

23 Tel.: +45 9940 8503; fax: +45 9814 1808.

24 Email address: phn@bio.aau.dk

25

1 **Summary**

2 Microbial communities are typically characterized by conditions of nutrient limitation so the availability of
3 the resources is likely a key factor in the niche differentiation across all species and in the regulation of the
4 community structure. In this study we have investigated whether four species exhibit any *in situ* short-term
5 changes in substrate uptake pattern when exposed to variations in substrate and growth conditions.
6 Microautoradiography was combined with fluorescence *in situ* hybridization to investigate *in situ* cell-
7 specific substrate uptake profiles of four probe-defined coexisting species in a wastewater treatment plant
8 with enhanced biological phosphorus removal. These were the filamentous “*Candidatus* Microthrix” and
9 *Caldilinea* (type 0803), the polyphosphate-accumulating organism “*Candidatus* Accumulibacter”, and the
10 denitrifying *Azoarcus*. The experimental conditions mimicked the conditions potentially encountered in
11 the respective environment (starvation, high/low substrate concentration, induction with specific substrates,
12 and single/multiple substrates). The results showed that each probe-defined species exhibited very distinct
13 and constant substrate uptake profile in time and space, which hardly changed under any of the conditions
14 tested. Such niche partitioning implies that a significant change in substrate composition will be reflected in
15 a changed community structure rather than the substrate uptake response from the different species.

16

17 **Introduction**

18 The enhanced biological phosphorus removal process (EBPR) is one of the most advanced wastewater
19 treatment configurations which removes carbon, nitrogen, and phosphorus from wastewater. A range of
20 modern molecular methods has revealed that the EBPR process is carried out by microbial communities
21 consisting primarily of uncultured microorganisms; that a low number of core microorganisms is critical to
22 the process, and that the communities in a wide range of EBPR plants seem to be stable at the genus level
23 (Nielsen *et al.*, 2010). Many of the important microbes have been identified and their main functions and
24 metabolic potential determined, making the EBPR process one of the best characterized ecosystems in
25 environmental biotechnology to date (Albertsen *et al.*, 2012, Nielsen *et al.*, 2012).

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25

Microbial community diversification is influenced by several environmental factors, such as phage predation (Šimek *et al.*, 2010) and substrate availability (Jasmin and Kassen, 2007). The EBPR communities typically live under dynamic conditions and nutrient limitation, thus the availability of the resources is likely to be the key factor in the regulation of populations within the community. Presence of specific substrates is important for the community structure in wastewater treatment systems, where very specialized heterotrophs have been found in several functional groups, such as denitrifiers (Thomsen *et al.*, 2007; Morgan-Sagastume *et al.*, 2008) and many filamentous bacteria (Nielsen *et al.*, 2009).

Pure culture studies, however, do not always support the concept that high substrate specificity is a general feature among heterotrophic bacteria. Batch and chemostat experiments with *Escherichia coli* and other pure cultures have shown, that under carbon/energy limited slow growth conditions, they change their substrate uptake profiles from consumption of only a few different substrates to utilization of many carbon sources (Ihssen and Egli, 2005; Liu *et al.*, 2005; Egli, 2010). It is hypothesized that expression of such behavior increases the chance of survival because, in natural environments, where the concentration of available carbon and energy sources is low, the expression of multiple transporters and catabolic enzymes gives the cell metabolic flexibility and makes the asset competitive.

The difference between studies carried out under *in situ* conditions and studies performed in pure culture raises the question of whether bacteria present in EBPR ecosystems and other complex ecosystems have a phenotypic plasticity to carry out short-term physiological change to be able to consume other substrates in response to changes in growth conditions such as starvation or variation in type or number of substrates available. Alternatively, such changes can only take place after genetic adaptation, typically after many generations, by adaptive mutations, gene loss, or horizontal gene transfer (Blount *et al.*, 2008; Shapiro *et al.*, 2009; Philippot *et al.*, 2010), allowing them to utilize the new resources and occupy new niches. It is

1 proposed that genetic microdiversity among closely related strains of a species in a natural environment
2 provides a pool of bacteria able to take over when their fitness is changing (Cohan, 2006, 2011; Ward,
3 2006). If this change is represented by the availability of another substrate, we should expect this to
4 become visible and of importance only when new clonal offspring become abundant after several
5 generations, unless the individual bacteria are immediately able to adjust to such a new situation and
6 change their substrate uptake patterns.

7

8 In this study, we hypothesized that (i) frequently observed functional bacterial groups in EBPR
9 communities (i.e., nitrifiers, denitrifiers, poly-phosphate accumulating bacteria, and filamentous bacteria)
10 are generally highly specialized in their substrate uptake profiles, i.e. have a clear niche partitioning, and
11 that (ii) these profiles are strongly preserved under exposure to short-term variations (few hours, but
12 repeatedly) such as substrate or oxygen concentrations, and type and number of substrates present in the
13 EBPR system. To reveal the short-term physiological changes, we have investigated the *in situ* substrate
14 uptake profiles of probe-defined species, by means of microautoradiography combined with fluorescence
15 *in situ* hybridization (MAR-FISH) under the above-mentioned conditions in a wastewater treatment plant
16 system. Four different probe-defined uncultured populations, here defined as “species”, that are typical for
17 EBPR communities, were selected: two filamentous species, the actinobacterial “*Candidatus* *Microthrix*”
18 (hereafter called *Microthrix*) and a *Caldilinea* (phylum *Chloroflexi*), and two microcolony-forming
19 bacteria, “*Candidatus* *Accumulibacter*” (hereafter called *Accumulibacter*), a key microorganism involved
20 in the phosphate removal, and *Azoarcus*, a common denitrifier. Previous studies have shown that these
21 bacteria form part of the core community in EBPR plants, they are abundant (each usually comprising
22 >2% of the entire community) and show clear differences in substrate uptake patterns (Nielsen *et al.*, 2010;
23 2012).

24

25

1 **Results**

2 To study the substrate uptake pattern of the probe-defined species, standard incubation conditions were
3 defined as 0.5-2 mM substrate and air saturation (defined as high substrate and high oxygen conditions).
4 These conditions are comparable to other ecophysiological studies carried out earlier and summarized by
5 Nielsen *et al.*, (2009, 2010). The uptake patterns of 11 substrates commonly encountered in wastewater
6 influent were evaluated by the MAR-FISH procedure and were in agreement with results from earlier
7 studies conducted under similar conditions over the past 10 years, giving a very consistent substrate uptake
8 pattern for each of the probe-defined species (Fig. 1 and Table S1).

9

10 Preferred position of Figure 1

11

12 The results show four specialized probe-defined species with different substrate uptake profiles. The
13 majority of the probe-defined filamentous *Microthrix* (> 80%) took up solely oleic acid with high uptake
14 activity (i.e., cells were heavily covered with silver grains) (Fig. 1A). Over 80% of filamentous *Caldilinea*
15 took up glucose with high activity. A fraction consumed oleic acid, mannose, and galactose as well,
16 indicating that *Caldilinea* was able to take up sugar compounds and oleic acid with glucose as the
17 preferred substrate (Fig. 1B). Microcolony-forming *Accumulibacter* and *Azoarcus* had more similar
18 uptake profiles and took up short-chain fatty acids such as acetate, propionate, and pyruvate (Figs. 1C and
19 1D). The pure culture of *E. coli* MG1655, included as a control organism typically used in laboratory
20 studies, exhibited very dynamic substrate uptake profile. Unlike the uptake profiles of probe-defined
21 EBPR species, the *E. coli* substrate uptake profile changed depending on the conditions applied. *E. coli*,
22 from exponential as well as stationary phase, took up a variety of organic compounds, such as short-chain
23 fatty acids, sugar compounds, amino acids, and glycerol (Figs. 1E and 1F and Fig. S1). The percentage of
24 active cells and the activity level of substrate uptake varied depending on the substrate.

25

1 **Substrate uptake at low substrate concentrations**

2 In order to investigate substrate uptake at low concentration, which is often encountered at some stages of
3 the wastewater treatment process, low substrate concentration was tested, here defined as 5-20 μM . Under
4 this condition, most *Microthrix* cells took up oleic acid, as in the case of high substrate concentration, but
5 the activity level was considerably lower. It was demonstrated by the lower number of silver grains formed
6 on top of individual cells (Fig. 1A), indicating that the substrate concentration was below the saturation
7 concentrations of the uptake system. *Caldilinea* utilized the same substrates under the low and high
8 substrate concentration conditions, although the number of filaments utilizing mannose and galactose
9 increased at the low substrate concentration (Fig. 1B). No difference was observed for *Accumulibacter* and
10 *Azoarcus* compared to the uptake at high substrate concentrations (Figs. 1C and 1D). Interestingly, the pure
11 culture of *E. coli* showed a weaker uptake capability of several substrates under the low substrate
12 concentration condition. This observation was most pronounced for stationary phase cells (Figs. 1E and
13 1F), and the change in *E. coli* substrate uptake pattern was much more pronounced than in the case of
14 microorganisms from natural communities.

15

16 **Effects of starvation**

17 Microorganisms in aeration tanks or clarifiers often experience starvation for 5-10 h, thus, in order to
18 investigate the effect of starvation on the substrate uptake profile, activated sludge (and *E. coli*) samples
19 were first deprived of an external carbon source for 12 h, combined with aeration, before the substrate
20 uptake profiles of the bacteria were investigated at high and low substrate and oxygen concentrations,
21 respectively. The substrate uptake patterns at high substrate and oxygen concentrations after starvation did
22 not change for *Microthrix*, *Accumulibacter* and *Azoarcus*. Minor differences concerning the percentage of
23 the population with a positive substrate uptake were observed for *Caldilinea*. In the case of incubation with
24 galactose at high substrate concentration, the number of positive cells increased for both high and low O_2
25 conditions while in the case of incubation with the same substrate under low substrate concentration, the

1 number of positive cells decreased under low O₂ conditions. In the case of incubation with glucose at low
2 substrate concentration for both high and low O₂ conditions, a decrease in glucose uptake activity was
3 observed. For *E. coli* the fraction of positive cells decreased for a few substrates. The differences were most
4 pronounced in the case of stationary phase *E. coli* incubated with high substrate concentration (for the cells
5 in exponential phase changes were observed in case of glycine (high substrate concentration) and acetate
6 and glycine (low substrate concentration); for the cells in stationary phase a decrease in the uptake of 6 out
7 of 9 substrates was observed when compared to standard high substrate concentration conditions, while no
8 changes were observed for low substrate concentration) (Figs. 1 E and F).

9

10 In separate series of experiments, the oxygen level was reduced from 20% to 1% (corresponding to
11 approx. 10 and 0.5 mg-oxygen l⁻¹, respectively) during starvation mimicking conditions commonly
12 encountered in activated sludge treatment plants. As some bacteria in activated sludge plants are
13 considered microaerophilic, e.g. *Microthrix* (Rossetti *et al.*, 2005), high oxygen concentration may be toxic
14 and prevent substrate uptake. However, no differences were observed in the uptake pattern of the probe-
15 defined species compared to the experiments carried out at air saturation (results not shown).

16

17 **Effects of induction**

18 Microorganisms in EBPR plants experience variations in type and concentrations of substrate during the
19 process due to variations in incoming wastewater content, e.g. from specific industries. In order to see
20 whether a prolonged exposure of the specialized bacteria to various specific substrates can induce a change
21 in uptake pattern, a series of experiments with only one substrate present in prolonged time (12 h), here
22 called induction experiments, was carried out. Samples were incubated for 12 h with selected substrates
23 under standard conditions (0.5-2 mM), followed by MAR incubations under the same substrate conditions
24 to observe the possible change in the substrate uptake profile. No changes were observed for *Microthrix*,
25 *Accumulibacter*, and *Azoarcus* (Figs. 1A, 1C, and 1D). For *Caldilinea*, slight changes were noticed. In the

1 case of glucose, the fraction of positive cells remained the same, although the observed activity level for
2 uptake decreased after induction with all substrates tested. In the case of mannose and galactose, the active
3 fraction of probe-defined bacteria increased with all substrates tested after the induction period (Fig. 1B).
4 For *E. coli* cells changes in the uptake pattern of galactose and both amino acids were observed in both
5 exponential and stationary phases (Figs. 1E and 1F). The uptake of glycine in exponential phase decreased
6 significantly after the leucine induction. Similarly, the active fraction of cells taking up leucine in stationary
7 phase decreased after the leucine induction. In the case of galactose, the uptake activity was higher in both
8 *E. coli* samples after induction with other sugars.

9

10 **Effects of multiple substrates**

11 As the presence of certain substrates is necessary for uptake of other substrates under *in situ* conditions
12 (Kong *et al.*, 2004; Thomsen *et al.*, 2007), a number of experiments were carried out to reveal such
13 potential dependencies. The substrate uptake pattern for all probe-defined species remained virtually
14 unchanged after co-incubations with chosen substrates, except for two important changes (Fig 1). Firstly,
15 *Microthrix* was able to utilize glycerol only when oleic acid was present as a co-substrate (Figs. 1A and
16 2A-2F). Secondly, *Accumulibacter* and *Azoarcus* were only capable of utilizing leucine when acetate was
17 present as a co-substrate (Figs. 1C and 1D and 2G-2I). Interestingly, *E. coli* exhibited changes in the uptake
18 pattern for a significant fraction of the substrates tested. These included changes in both the percentage of
19 positive cells and the activity level of substrate uptake. For exponential *E. coli* cells a decrease in the
20 fraction of positive cells could be observed in the case of 6 substrates tested. The response of stationary *E.*
21 *coli* cells was more diverse as both decrease as well as increase in the fraction of metabolically active cells
22 could be observed for different substrates. Such a heterogeneous behavior could be noticed in the case of
23 glycerol as the presence of different co-substrates induced different metabolic responses.

24

25 Preferred position of Figure 2

1

2 **Discussion**

3 This study shows that microorganisms growing under dynamic conditions in EBPR plants were strongly
4 specialized in their choice of substrate and that any changes in their *in situ* substrate uptake profiles were
5 very limited when exposed to short-term variations in substrate and growth conditions, typical for the
6 environment investigated. Different types of substrates, concentrations, starvation as well as induction with
7 selected substrates did not induce the uptake of other substrates. Only in the presence of multiple substrates
8 three cases were observed where substrates not consumed under standard conditions were co-utilized in
9 the presence of a second substrate. Our observations suggest that the phenotypic plasticity of the core
10 species in the EBPR plant was low as originally hypothesised, and thus a possible adaptation could only
11 occur through an evolutionary response and the acquisition of mutation that allow access to a novel
12 substrate. These results were distinctly different from the patterns exhibited by *E. coli*, where the ability to
13 utilize individual substrates depended on the environmental conditions. These pure-culture observations
14 are consistent with other literature reports on the state of preparedness of bacteria grown in pure cultures to
15 utilize multiple substrates simultaneously for low (Ihssen and Egli, 2005; Liu *et al.*, 2005) as well as high
16 (Egli, 1995) substrate concentration conditions.

17

18 The four probe-defined species had distinctly different substrate uptake profiles, showing the existence of
19 different stable ecological niches in the wastewater system investigated. This is not only the case for the
20 four species investigated here, but also for most of the approx. 30 core species present in the EBPR system,
21 where the substrate profiles have been stable for several years in the specific probe-defined populations
22 (Nielsen *et al.*, 2010; 2012). As an example, the substrate uptake profile of *Microthrix* was investigated
23 regularly over the past 15 years in many different plants, always with the same result (Andreasen and
24 Nielsen, 1997; 2000, Nielsen *et al.*, 2002, and this study). These results show that the species present in the
25 EBPR plants are very stable in substrate uptake profile. Although genetic adaptation potentially will occur

1 over many generations in the presence of a multitude of related niches and perfect conditions for acquiring
2 adaptive genes, such offspring did not seem to be competitive, nor able to establish themselves. This is
3 important for past and future *in situ* investigations of substrate uptake profiles in EBPR plants, in other
4 engineered systems, and potentially in other natural ecosystems.

5

6 Little is known about the microdiversity among activated sludge bacteria, but recently, significant
7 microdiversity was shown in *Accumulibacter* from full-scale treatment plants (Albertsen *et al.*, 2012). The
8 differences between the genome of the activated sludge *Accumulibacter* and the only existing reference
9 genome (Garcia Martin *et al.*, 2006) are primarily related to the genes involved in the production of
10 extracellular polymers and different virus-related genes and are not linked to the central metabolic
11 pathways. Presently, no other genomes exist with a sufficient similarity to the core species in the EBPR
12 plants to perform more detailed comparisons (Albertsen *et al.*, 2012). Wastewater treatment plants
13 certainly expose microbial populations to conditions that could favor periodic selections or “sweeps”
14 (Cohan and Koeppel, 2008; Fraser *et al.*, 2009), but no new strains with changed substrate uptake pattern
15 seemed to evolve.

16

17 In the case of *Accumulibacter*, *Azoarcus*, and *Microthrix*, the presence of a second substrate induced the
18 uptake of substrates that were not taken up when present as sole substrates. *Accumulibacter* and *Azoarcus*
19 took up leucine only when acetate was present, as reported previously (Kong *et al.*, 2004; Thomsen *et al.*,
20 2007). It was hypothesized that leucine could potentially be used as a nitrogen source, but the detailed
21 mechanism for its uptake and use is not clear. A notable fraction of *Microthrix* was able to consume
22 glycerol but only when oleic acid was present. As *Microthrix* produce extracellular lipases that degrade
23 triglycerides into long chain fatty acids and glycerol (Nielsen *et al.*, 2002), it seems rational that they can
24 consume both types of substrate, but the exact mechanism is not known, and no genomic information is
25 available.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25

Although it is not known how common the simultaneous uptake of several substrates by probe-defined species in activated sludge is, this physiological trait presumably gives a competitive advantage, particularly in the dynamic complex communities, in which several organic compounds might be available in the local environment. However, the results show that the exact substrate uptake profile for a certain species cannot always be based on single-substrate experiments only.

Caldilinea species remained specific towards four substrates throughout all the conditions applied, although slight changes in sugar uptake activity were observed at low substrate concentration, starvation, and induction conditions. No K_m values are reported in the literature for *Caldilinea*; thus it is not clear if substrate affinity or other factors are responsible for the observed changes in the uptake of monosaccharides. In some cases, the probe-defined *Caldilinea* showed increased galactose and mannose uptake while the activity of glucose uptake was lowered. However, the ability to modify individual substrate uptake was highly dependent on the type of conditions applied. It seems possible that the *Caldilinea* population was able to respond to unfavorable conditions (starvation, diminished substrate availability) as well as to induction with all three monosaccharides tested. It was not an uptake of new substrates, but a difference in the relative uptake rate of the substrates.

The specificity of FISH probes applied in this study was high, typically covering a single or few species in a genus as characterized by 16S rRNA gene phylogeny. Three out of four probe-defined species showed very consistent substrate profiles for all bacteria targeted by the probes, indicating little diversity in terms of substrate specificity. Only the *Caldilinea* group exhibited a slightly different substrate uptake pattern for a subpopulation presumably due to the presence of more than one phenotype or due to clonal heterogeneity (Elowitz *et al.* 2002).

1 Trying to extrapolate potential substrate capabilities of the uncultured probe-defined *Azoarcus* and
2 *Caldilinea* based on their isolates is clearly not possible. *Azoarcus* spp. can grow on glucose and ethanol in
3 pure cultures (Zhou *et al.*, 1995), but was not able to consume these substrates under *in situ* conditions.
4 Furthermore, pure culture of filamentous *Caldilinea* sp. isolated from wastewater treatment plants could
5 utilize acetate and pyruvate (Yoon *et al.*, 2010), in contrast to *in situ* studies. Also other closely related
6 species isolated from plant roots or rivers showed very different substrate uptake capabilities (Misko and
7 Germida, 2002; Freese *et al.*, 2010). The conclusion from our results indicates that, although great diversity
8 may evolve around each species present in the EBPR ecosystems, the strong competition in the
9 community only allows a single or very few strains of this species with conserved substrate uptake profile
10 to multiply and become abundant. It is well known from pure culture studies that just a few hundred
11 generations may be enough for new subpopulations to evolve and adapt to new distinct substrates (Jasmin
12 and Kassen, 2007; Blount *et al.*, 2008; Lee *et al.*, 2009), but under *in situ* conditions in EBPR plants, they
13 appear not to have the enhanced fitness to become dominant. However, they may be present in very low
14 abundance not observed by the FISH method applied, that could potentially be seen as a range of closely
15 related strains or species with different metabolic potentials for substrate utilization. This underlines the
16 need for obtaining reference genomes or isolates from the actual dominant members of EBPR
17 communities, and not from other closely related strains, for predicting their function in the ecosystem by
18 metabolic reconstruction.

19

20 Our results suggest limitations concerning the use of leucine as an indicator of protein synthesis
21 (heterotrophic production), bacterial growth, or activity in mixed ecosystems (Kirchman *et al.*, 1985;
22 Alonso-Sáez *et al.*, 2010). We showed that leucine was not incorporated by any of the four probe-defined
23 bacteria under the conditions tested, with two exceptions (uptake of leucine by *Accumulibacter* and
24 *Azoarcus* when acetate was present). Leucine uptake in *E. coli* was strongly dependent on the conditions
25 applied. This is in agreement with investigations of an estuarine community, where the incorporation of

1 leucine was not suitable to determine/estimate the cell activity (Mayali *et al.*, 2011). These results show that
2 leucine cannot be used as a general measure for microbial activity in natural ecosystems without careful
3 controls.

4

5 The study clearly demonstrates that bacteria present in the EBPR ecosystem, forming a stable core
6 community across many plants, were specialized in uptake of specific substrates, although with some
7 substrate overlap. It is, however, interesting that we have not been able to find a broad generalist able to
8 consume the majority of the substrates tested in this or previous studies (summarized by Nielsen *et al.*,
9 2010; 2012). The niche overlap observed for most substrates stresses the fact that other physiological
10 factors are also important for defining the niches for various species under natural dynamic conditions,
11 such as substrate affinity, activity under different electron acceptor conditions, and the ability to store
12 substrates. The implication of such niche partitioning is also that a significant change in the incoming
13 substrate composition will immediately be reflected in a changed community structure and not in
14 alterations in the substrate uptake response from the individual species. This is in agreement with the
15 observations of community shift in full-scale plants with external carbon addition for improved
16 denitrification (Hagman *et al.*, 2008). This observation further underlines that neutral models that do not
17 take such niches into account when describing factors shaping communities (Ofiteru *et al.*, 2010) cannot
18 reliably describe all microbial ecosystems.

19

20 Interestingly, the results indicate that, with some precautions, it is possible to make community-wide
21 predictions of microbial substrate networks in the EBPR system and other engineered microbial
22 ecosystems, if relevant reference genomes are known and the system is calibrated by proper measurements
23 of the substrate uptake capabilities under the actual *in situ* conditions. Such predictions would be difficult if
24 extensive, short-term, physiological adaptations took place as illustrated by the studies of *E. coli*. If *E. coli*
25 substrate uptake should be predicted under different environmental conditions, based on its genomic

1 potential only, our results show that this would be difficult, unless coupled with analyses of expressed
2 genes, proteins or other data showing the actual *in situ* activity. To what extent this is also the case for
3 bacteria in more oligotrophic natural environments remains to be investigated.

4

5 **Experimental procedures**

6 **Activated sludge samples**

7 Activated sludge was collected from the Aalborg West full-scale WWTP. The plant is designed for a
8 population equivalent of 300,000 and performs biological nitrogen and phosphorus removal by the
9 alternating BioDenipho process (Seviour *et al.*, 2010). The WWTP has previously been described in detail
10 (Nielsen *et al.*, 2002). All samples were collected from the aeration tanks on the same day as the
11 experiments were conducted. Before the experiments, activated sludge was gently homogenized and
12 aerated for 1-2 h to reduce possible background substrate in the samples.

13

14 **Bacterial strain and growth conditions**

15 Wild-type *E. coli* K-12 MG1655 was used as a control in all experiments. Complex medium used for *E.*
16 *coli* cultivation contained 10 g l⁻¹ tryptone, 5 g l⁻¹ yeast extract, 12.8 g l⁻¹ Na₂HPO₄·2H₂O, 3 g l⁻¹ KH₂PO₄,
17 1.77 g l⁻¹ (NH₄)₂SO₄, 130 mg l⁻¹ MgCl₂·6H₂O, 80 mg l⁻¹ CaCO₃, 77 mg l⁻¹ FeCl₃·6H₂O, 11 mg l⁻¹
18 MnCl₂·4H₂O, 1.5 mg l⁻¹ CuSO₄·5H₂O, 1.3 mg l⁻¹ CoCl₂·6H₂O, 4 mg l⁻¹ ZnO, 1.2 mg l⁻¹ H₃BO₃, 8.5 mg l⁻¹
19 NaMoO₄, 790 mg l⁻¹ EDTA Na₄·2H₂O as described by Ihssen and Egli (2004, 2005). All salts were
20 added as concentrated solutions to the medium. Bacteria were grown in Erlenmeyer flasks at 37°C on
21 rotary shaker (at 150 rpm) and collected during the exponential (2.5 h, OD_{546nm} = 1.3 ± 0.4) and stationary
22 (24 h, OD_{546nm} = 7.8 ± 0.5) phases. Each sample was washed 3 times with the mineral medium to remove
23 carbon source from the medium and resuspended in the mineral medium preheated to 37°C. The mineral
24 medium was prepared in the same way as the complex medium above, but without tryptone and yeast
25 extract.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25

Incubation procedures

Multiple experiments were conducted to study the effect of high or low substrate concentration, starvation, induction with other substrates, and multiple substrates, on the substrate uptake specificity of investigated bacterial species. Starvation experiments were performed both under high oxygen concentration (20%, corresponding to normal air saturation, 10 mg Γ^{-1}) and low oxygen concentration (1%, 0.5 mg Γ^{-1}). Low oxygen concentration in the headspace was achieved by injecting appropriate volume of air into the serum bottles after repeated (three times) evacuation of the headspace with high purity N_2 (99.999%). Each starvation condition tested was followed by incubation with both high and low substrate concentrations. Inductions and multiple substrate incubations were followed by incubation under the high oxygen and high substrate condition. An overview of experimental procedures is shown in Table S2 and S3.

Organic substrates labeled with radioisotopes used for MAR incubations are described in SI Methods. All unlabeled organic compounds were prepared as concentrated stock solutions (20 or 80 mM) with the pH adjusted to 7.5. In all the incubations, the ratio between labeled and unlabeled substrates as well as the ratio between the substrate and biomass concentration (specific activity, 5 $\mu\text{Ci mgSS}^{-1}$) were identical to allow the substrate uptake to be compared between different experiments. The conditions were adjusted in order to ensure that the substrates were not depleted during the incubations. All incubations were conducted for 2 h, aerobically in the dark at 20°C (activated sludge) or 37°C (*E. coli*). The serum bottles were sealed with gas-tight rubber stoppers and gently shaken at 150 rpm during the incubation. Biological replicates were performed. For each biological replicate two technical replicates were prepared. After MAR incubation, all samples were fixed for 3 h (activated sludge) or 1 h (*E. coli*) either with 4% [w/v] paraformaldehyde (Gram-negative bacteria, i.e., *Caldilinea*, *Accumulibacter*, *Azoarcus*, and *E. coli*) or 50% [v/v] ethanol (Gram-positive bacteria, i.e., *Microthrix*).

1 **High and low substrate concentration conditions.** Activated sludge was diluted with filtered effluent
2 water from the same treatment plant to a final SS (dry matter) concentration of 1 g-SS Γ^{-1} for experiments
3 at the high substrate concentration condition (0.5-2 mM). At low concentrations (5-20 μ M) the sludge was
4 diluted to 0.01 g-SS Γ^{-1} to ensure a low overall substrate removal rate and that substrates were not depleted.
5 *E. coli* was diluted with mineral media to a final dry matter concentration of 0.1 g-SS Γ^{-1} and 0.001 g-SS Γ^{-1}
6 for high and low substrate concentration conditions, respectively. All samples were incubated with
7 individual radioactively labeled and unlabeled organic substrates. The concentrations of all substrates used
8 are listed in Table S2. For the high substrate concentration condition, 2 ml of diluted sample was
9 transferred to 10-ml serum bottle. For the low concentration condition, 20 ml of diluted sample was
10 transferred to 110-ml serum bottle.

11

12 **Starvation conditions.** Undiluted activated sludge (30 ml, 3.5 g-SS Γ^{-1}) and *E. coli* were transferred to 300-
13 ml serum bottles and starved for 12 h under aerobic conditions. Serum bottles were incubated on rotary
14 shaker (150 rpm) at 20°C (activated sludge) and 37°C (*E. coli*). After starvation, the sludge and *E. coli*
15 were incubated under both high and low substrate and oxygen conditions, according to the procedure
16 described above.

17

18 **Induction conditions.** Four induction experiments were performed (Table S2). 50 ml of diluted-sludge
19 (0.2 g-SS Γ^{-1}) and *E. coli* (0.1 g-SS Γ^{-1}) were transferred to 300-ml serum bottles. Organic substrates (3
20 mM) were added, and cells were induced for 12 h under aerobic conditions. Serum bottles were incubated
21 on a rotary shaker (150 rpm) at either 20°C (activated sludge) or 37°C (*E. coli*). After induction, the sludge
22 and *E. coli* were washed three times with filtered tap water and incubated with chosen labeled substrates
23 (Table S2) under the high substrate and oxygen concentration conditions, according to the procedure
24 described above.

25

1 **Multiple substrate conditions.** Diluted-sludge (2 ml, 1 g-SS Γ^{-1}) and *E. coli* (0.1 g-SS Γ^{-1}) were transferred
2 to 10-ml serum bottles and incubated with one labeled substrate and one unlabeled organic substrate (Table
3 S3).

4 5 **Fluorescence *in situ* hybridization and DAPI staining**

6 Fixed activated sludge and *E. coli* cells were washed three times with filtered tap water and then
7 resuspended in a 1:1 volume ratio of phosphate-buffered saline (PBS) and 99.9% ethanol. Fixed sludge
8 was spread onto gelatin-coated glass slides, air-dried, and dehydrated consecutively in 50%, 80%, and
9 99.9% [v/v] ethanol. FISH was carried out according to Amann (1995). The following oligonucleotide
10 probes were used for FISH: EUBmix [equimolar concentrations of EUB338 (Amann *et al.*, 1990),
11 EUB338II and EUB338III (Daims *et al.*, 1999)] targeting most of the *Bacteria*, MPA60 (Erhart *et al.*,
12 1997) targeting *Microthrix* (class *Actinobacteria*, phylum *Actinobacteria*), T0803-0654 (Kragelund *et al.*,
13 2011) targeting “type 0803” within the *Caldilinea* genus (class *Caldilineae*, phylum *Chloroflexi*), Acc-I-
14 444 (Flowers *et al.*, 2009) targeting *Accumulibacter* clade I (class *Betaproteobacteria*), and Azo644 (Hess
15 *et al.*, 1997) targeting *Azoarcus* (order *Rhodocyclales*, class *Betaproteobacteria*), EUBmix probe was
16 labeled with 5,(6)-carboxyfluorescein-*N*-hydroxysuccinimide ester (FLUOS), all remaining probes were
17 labeled with sulfoindocyanine dye (Cy3). Fixed *E. coli* cells were spread onto gelatin-coated glass slides,
18 airdried and stained with DAPI solution for 5 min in the dark (Porter and Feig, 1980) after the
19 microautoradiographic (MAR) procedure.

20

21 **Microautoradiography**

22 The combination of MAR and FISH was carried out as described by Nielsen and Nielsen (2005). Slides
23 with activated sludge (after FISH procedure) and *E. coli* cells were coated with LM-1 emulsion (GE
24 Healthcare UK Ltd., Little Chalfont, United Kingdom), exposed in the dark for 3 and 10 days and then
25 developed with Kodak D-19 developer, as described by Nielsen and Nielsen (2005). In this report, the

1 results after 10 days exposure are discussed (Fig. 1). However, it must be noted that the substrate uptake
2 patterns observed after 3 and 10 days exposure were similar (data not shown).

3

4 **Microscopy**

5 An epifluorescence microscope (Axioskop 2 Plus, Zeiss, Oberkochen, Germany) equipped with a charge-
6 couple device (CCD) camera (CoolSNAP HQ, Photometrics, Oberkochen, Germany) was used to
7 examine all the slides. Silver grain formation on probe-defined or DAPI-stained bacteria was observed
8 with light microscopy. MAR-positive and MAR-negative cells were assessed visually by comparison of
9 the number of silver grains developed on top of the cells to the background level.

10

11

12 **Acknowledgements**

13 This research was partially supported by the JSPS Institutional Program for Young Researcher Overseas
14 Visits (T. Kindaichi), by research foundation from Public Association for Construction Service Chugoku
15 region (T. Kindaichi), by FøSu (M. Nierychlo and P.H. Nielsen), and by Aalborg University.

16

17

18 **References**

19 Albertsen, M., Hansen, L.B., Saunders, A.M., Nielsen, P.H., and Nielsen, K.L. (2012) A metagenome of a
20 full-scale microbial community carrying out enhanced biological phosphorus removal. *ISME J* **6**:
21 1094–1106.

22 Alonso-Sáez, L., Galand, P.E., Casamayor, E.O., and Pedrós-Alió, C. (2010) High bicarbonate
23 assimilation in the dark by Arctic bacteria. *ISME J* **4**: 1581–1590.

- 1 Amann, R.I., Binder, B.J., Olson, R.J., Chisholm, S.W., Devereux, R., and Stahl, D.A. (1990) Combination
2 of 16S rRNA-targeted oligonucleotide probes with flow cytometry for analyzing mixed microbial
3 populations. *Appl Environ Microbiol* **56**: 1919–1925.
- 4 Amann, R.I. (1995) *In situ* identification of micro-organisms by whole cell hybridization with rRNA-
5 targeted nucleic acid probes. In: *Molecular Microbial Ecology Manual*. Akkermans, A.D.L., Van Elsas,
6 J.D., and De Bruijn, F.J. (eds). London, UK: Kluwer Academic Publications, pp. MEM-3.3.6/1–
7 MEM-3.3.6/15.
- 8 Andreasen, K., and Nielsen, P.H. (1997) Application of microautoradiography to the study of substrate
9 uptake by filamentous microorganisms in activated sludge. *Appl Environ Microbiol* **63**: 3662–3668.
- 10 Andreasen, K., and Nielsen, P.H. (2000) Growth of *Microthrix parvicella* in nutrient removal activated
11 sludge plants, studies of *in situ* physiology. *Water Res* **34**: 1559–1569.
- 12 Blount, Z.D., Borland, C.Z., and Lenski, R.E. (2008) Historical contingency and the evolution of a key
13 innovation in an experimental population of *Escherichia coli*. *Proc Natl Acad Sci USA* **105**:
14 7899–7906.
- 15 Cohan, F.M. (2006) Towards a conceptual and operational union of bacterial systematics, ecology and
16 evolution. *Philos Trans R Soc Lond B Bio Sci* **361**: 1985–1996.
- 17 Cohan, F. M. (2011) Are species cohesive – a view from bacteriology. “Population Genetics of Bacteria: a
18 Tribute to Thomas S. Whittam”. Ed. S. Walk and P. Feng. Washington: American Society for
19 Microbiology Press, 2011. 43-65.
- 20 Cohan, F.M., and Koeppl, A.F. (2008) The origins of ecological diversity in prokaryotes. *Curr Biol* **18**:
21 R1024–R1034.
- 22 Daims, H., Bruhl, A., Amann, R., Schleifer, K.H., and Wagner, M. (1999) The domain-specific probe
23 EUB338 is insufficient for the detection of all bacteria: development and evaluation of a more
24 comprehensive probe set. *Syst Appl Microbiol* **22**: 434–444.

- 1 Egli, T. (1995) The ecological and physiological significance of the growth of heterotrophic
2 microorganisms with mixtures of substrates. *Adv Microb Ecol* **14**: 305–386.
- 3 Egli, T. (2010) How to live at very low substrate concentration. *Water Res* **44**: 4826–4837.
- 4 Elowitz M.B., Leine A.J., Siggia E.D., and Swain P.S. (2002) Stochastic gene expression in a single cell.
5 *Science* **297**: 1183–1186.
- 6 Erhart, R., Bradford, D., Seviour, R.J., Amann, R., and Blackall, L.L. (1997) Development and use of
7 fluorescent in situ hybridization probes for the detection and identification of ‘*Microthrix parvicella*’ in
8 activated sludge. *Syst Appl Microbiol* **20**: 310–318.
- 9 Flowers, J.J., He, S., Yilmaz, S., Noguera, D.R., and McMahon, K.D. (2009) Denitrification capabilities of
10 two biological phosphorus removal sludges dominated by different ‘*Candidatus Accumulibacter*’
11 clades. *Environ Microbiol Rep* **6**: 583–588.
- 12 Fraser, C., Alm, E.J., Polz, M.F., Spratt, B.G., and Hanage, W.P. (2009) The bacterial species challenge:
13 Making sense of genetic and ecological diversity. *Science* **323**: 741–746.
- 14 Freese, H.M., Karsten, U., and Schumann, R. (2010) Bacterial abundance, activity, and viability in the
15 eutrophic river Warnow, Northeast Germany. *Microb Ecol* **51**: 117–127.
- 16 García Martín, H., Ivanova, N., Kunin, V., Warnecke, F., Barry, K.W., and McHardy, A.C. *et al.* (2006)
17 Metagenomic analysis of two enhanced biological phosphorus removal (EBPR) sludge communities.
18 *Nat Biotechnol* **24**: 1263–1269.
- 19 Hagman, H., Nielsen, J.L., Nielsen, P.H., and Jansen, J.C. (2008) Mixed carbon sources for nitrate
20 reduction in activated sludge - identification of bacteria and process activity studies. *Water Res* **42**: 1539
21 –1546.
- 22 Hess, A., Zarda, B., Hahn, D., Häner, A., Stax, D., Höhener, P., and Zeyer, J. (1997) In situ analysis of
23 denitrifying toluene- and *m*-xylene-degrading bacteria in a diesel fuel-contaminated laboratory aquifer
24 column. *Appl Environ Microbiol* **63**: 2136–2141.

- 1 Ihssen, J., and Egli, T. (2004) Specific growth rate and not cell density controls the general stress response
2 in *Escherichia coli*. *Microbiology* **150**: 1637–1648.
- 3 Ihssen, J., and Egli, T. (2005) Global physiological analysis of carbon- and energy-limited growing
4 *Escherichia coli* confirms a high degree of catabolic flexibility and preparedness for mixed substrate
5 utilization. *Environ Microbiol* **7**: 1568–1581.
- 6 Jasmin, J.N., and Kassen, R. (2007) On the experimental evolution of specialization and diversity in
7 heterogeneous environments. *Ecol lett* **10**: 272–281.
- 8 Kirchman, D., K'nees, E., and Hodson, R. (1985) Leucine incorporation and its potential as a measure of
9 protein synthesis by bacteria in natural aquatic systems. *Appl Environ Microbiol* **49**: 599–607.
- 10 Kong, Y., Nielsen, J.L., and Nielsen, P.H. (2004) Microautoradiographic study of *Rhodocyclus*-related
11 poly-P accumulating bacteria in full-scale enhanced biological phosphorus removal plants. *Appl*
12 *Environ Microbiol* **70**: 5383–5390.
- 13 Kragelund, C., Thomsen, T.R., Mielczarek, A.T., and Nielsen, P.H. (2011) Eikelboom's morphotype 0803
14 in activated sludge belongs to the genus *Caldilinea* in the phylum *Chloroflexi*. *FEMS Microbiol Ecol*
15 **76**: 451–462.
- 16 Lee, M.C., Chou, H.H., and Marx, C.J. (2009) Asymmetric, bimodal trade-offs during adaptation of
17 *Methylobacterium* to distinct growth substrates. *Evolution* **63**: 2816–2830.
- 18 Liu, M., Durfee, T., Cabrera, J.E., Zhao, K., Jin, D.J., and Blattner, F.R. (2005) Global transcriptional
19 programs reveal a carbon source foraging strategy by *Escherichia coli*. *J Biol Chem* **280**:
20 15921–15927.
- 21 Mayali, X., Weber, P.K., Brodie, E.L., Mabery, S., Hoerich, P.D., and Pett-Ridge, J. (2011) High-
22 throughput isotopic analysis of RNA microarrays to quantify microbial resource use. *ISME J* **6**:
23 1210–1221.
- 24 Misko, A.L., and Germida, J.L. (2002) Taxonomic and functional diversity of pseudomonads isolated
25 from the roots of field-grown canola. *FEMS Microbiol Ecol* **42**: 399–407.

- 1 Morgan-Sagastume, F., Nielsen, J.L., and Nielsen, P.H. (2008) Substrate-dependent denitrification of
2 abundant probe-defined denitrifying bacteria in activated sludge. *FEMS Microbiol Ecol* **66**: 447–461.
- 3 Nielsen, P.H., Roslev, P., Dueholm, T., and Nielsen, J.L. (2002) *Microthrix parvicella*, a specialized lipid
4 consumer in anaerobic – aerobic activated sludge plants. *Water Sci Technol* **46**: 73–80.
- 5 Nielsen, J.L., and Nielsen, P.H. (2005) Advances in microscopy: microautoradiography of single cells.
6 *Methods Enzymol* **397**: 237–256.
- 7 Nielsen, P.H., Kragelund, C., Seviour, R.J., and Nielsen, J.L. (2009) Identity and ecophysiology of
8 filamentous bacteria in activated sludge. *FEMS Microbiol Rev* **33**: 969–998.
- 9 Nielsen, P.H., Mielczarek, A.T., Kragelund, C., Nielsen, J.L., Saunders, A.M., and Kong, Y. (2010) A
10 conceptual ecosystem model of microbial communities in enhanced biological phosphorus removal
11 plants. *Water Res* **44**: 5070–5088.
- 12 Nielsen, P.H., Saunders, A.M., Hansen, A.A., Larsen, P., and Nielsen, J.L. (2012) Microbial communities
13 involved in enhanced biological phosphorus removal from wastewater - a model system in
14 environmental biotechnology. *Curr Opin Biotechnol* **23**: 452–459.
- 15 Ofiteru, I.D., Lunn, M., Curtis, T.P., Wells, G.F., Criddle, C.S., Francis, C.A., and Sloan, W.T. (2010)
16 Combined niche and neutral effects in a microbial wastewater treatment community. *Proc Natl Acad*
17 *Sci USA* **107**: 15345–15350.
- 18 Philippot, L., Andersson, S.G., Battin, T.J., Prosser, J.I., Schimel, J.P., Whitman, W.B., and Hallin, S.
19 (2010) The ecological coherence of high bacterial taxonomic ranks. *Nat Rev Microbiol* **8**: 523–529.
- 20 Porter, K.G., and Feig, Y.S. (1980) The use of DAPI for identifying and counting aquatic microflora.
21 *Limnol Oceanogr* **25**: 943–948.
- 22 Rossetti, S., Tomei M.C., Nielsen P.H., and Tandoi V. (2005) "*Microthrix parvicella*", a filamentous
23 bacterium causing bulking and foaming in activated sludge systems: a review of current knowledge.
24 *FEMS Microbiol Rev* **29**: 49-64.

1 Seviour, R.J., Lindrea, K.C., and Oehmen, A. (2010) The activated sludge process. In: *Microbial Ecology*
2 *of Activated Sludge*. Seviour, R.J., and Nielsen, P.H. (eds) London, UK: IWA publishing, pp. 57–94.

3 Shapiro, B.J., David, L.A., Friedman, J., and Alm, E.J. (2009) Looking for Darwin’s footprints in the
4 microbial world. *Trends Microbiol* **17**: 196–204.

5 Šimek, K., Kasalický, V., Hornák, K., Hahn, M.W., and Weinbauer, M.G. (2010) Assessing niche
6 separation among coexisting *Limnohabitans* strains through interactions with a competitor, viruses, and
7 a bacterivore. *Appl Environ Microbiol* **76**: 1406–1416.

8 Thomsen, T.R., Kong, Y., and Nielsen, P.H. (2007) Ecophysiology of dominant denitrifying bacteria in
9 activated sludge. *FEMS Microbiol Ecol* **60**: 370–382.

10 Ward, N. (2006) New directions and interactions in metagenomics research. *FEMS Microbiol Ecol* **55**:
11 331–338.

12 Yoon, D.N., Park, S.J., Kim, S.J., Jeon, C.O., Chae, J.C., and Rhee, S.K. (2010) Isolation, characterization,
13 and abundance of filamentous members of *Caldilineae* in activated sludge. *J Microbiol* **48**: 275–283.

14 Zhou, J., Fries, M.R., Chee-Sanford, J.C., and Tiedje, J.M. (1995) Phylogenetic analyses of a new group of
15 denitrifiers capable of anaerobic growth on toluene and description of *Azoarcus toluolyticus* sp. nov. *Int J*
16 *Syst Bacteriol* **45**: 500–506.

17
18
19
20
21
22
23
24
25

1 **Figure legends**

2 **Figure 1**

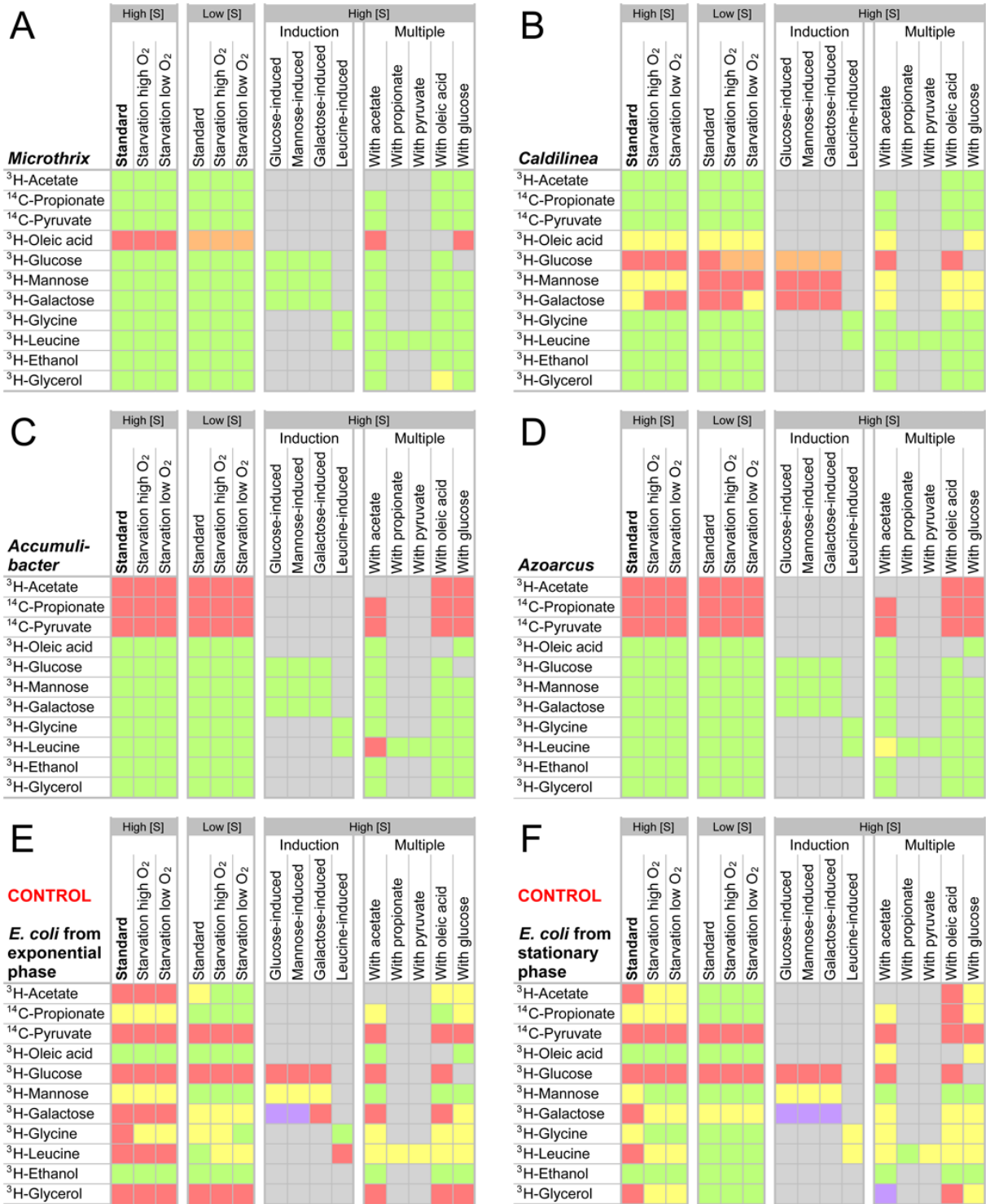
3 Summary of the substrate-uptake specificity for four probe-defined bacterial functional groups (*Microthrix*,
4 *Caldilinea*, *Accumulibacter*, *Azoarcus*) from Aalborg West activated sludge (A-D) and pure culture *E. coli*
5 MG 1655 from exponential and stationary phase (E and F, respectively), based on MAR experiments with
6 11 different labeled substrates performed under 15 different aerobic conditions. High [S] and low [S] in the
7 header mean high or low concentration of substrate, 0.5-2 mM and 5-20 μ M, respectively. Standard
8 conditions: 0.5-2 mM substrate and air saturation. High oxygen: air saturation. Low oxygen: 1% in
9 headspace, 0.5 mg-oxygen Γ^{-1} . Starvation: 12 h pre-incubation without any substrates. Induction: 12 h pre-
10 incubation with organic substrate (3 mM). Multiple substrates: incubation with two different (one labeled
11 and one unlabeled) substrates. The colors indicate fractions of active population and uptake activity.
12 Green: no substrate uptake, most cells negative (> 80% of total cells); yellow: some positive cells (20-80%
13 of total cells) exhibiting substrate uptake; red: most of the cells positive (> 80% of total cells) and highly
14 covered with silver grains (i.e., high activity); orange shows a decrease in uptake activity compared to red,
15 most of the cells positive (> 80% of total cells) with less silver grains per cell; purple shows an increase in
16 uptake activity compared to red, most of the cells positive (> 80% of total cells) with more silver grains per
17 cell. Gray: not determined.

18

19 **Figure 2**

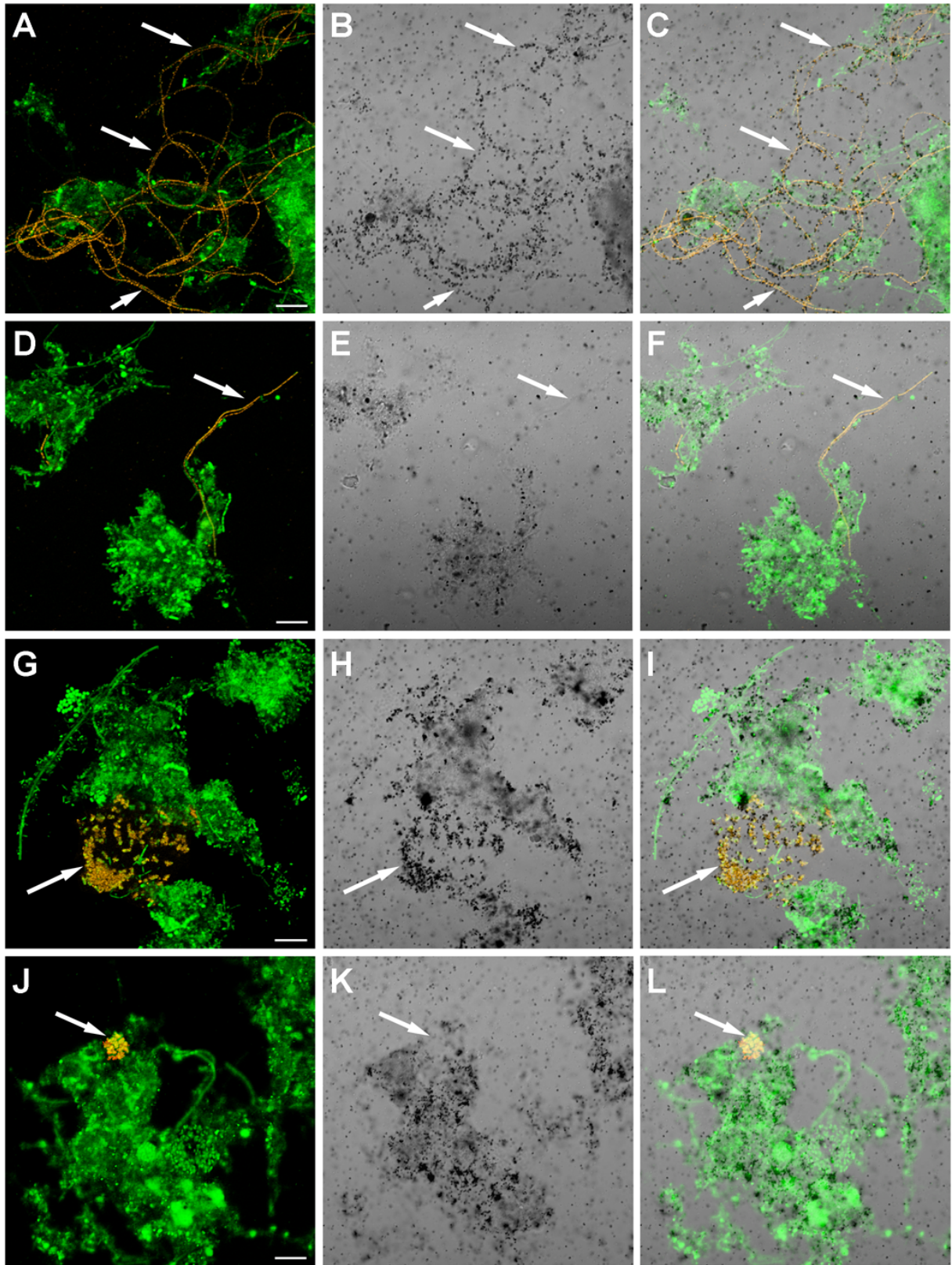
20 MAR-FISH micrographs of Aalborg West WWTP activated sludge sample incubated with labeled
21 glycerol under the multiple substrates (oleic acid + glycerol) condition (A-C), with labeled glycerol under
22 the high substrate concentration (standard) condition (D-F), with labeled leucine under the multiple
23 substrates (acetate + leucine) condition (G-I), and with labeled leucine under the high substrate
24 concentration condition (J-L). *In situ* hybridizations were performed with a combination of the FLUOS-
25 labeled mixed EUB338 probes (green), Cy3-labeled probes MPA60 (A and D), and Cy3-labeled probes

1 Acc-I-444 (G and J). *Microthrix* cells (A and D) and *Accumulibacter* cells (G and J) appear yellow, other
2 bacteria appear green. Panel B, E, H, and K are bright-field MAR images, corresponding to the FISH
3 images (A, D, G, and J, respectively). Panel C, F, I, and L are overlay images showing uptake of glycerol in
4 presence of oleic acid by *Microthrix* cells (C), lack of glycerol uptake by *Microthrix* cells when present as a
5 single substrate (F), uptake of leucine in presence of acetate by *Accumulibacter* cells (I), lack of leucine
6 uptake by *Accumulibacter* cells when present as a single substrate (L). The bars represent 10 μm . Arrows
7 indicate MAR-positive or MAR-negative cells.



1

2 Figure 1



1

2 Figure 2

1 **Supporting Information**

2 **High and stable substrate specificities of microorganisms in enhanced biological phosphorus**

3 **removal plants**

4

5 Tomonori Kindaichi^{a,b}, Marta Nierychlo^b, Caroline Kragelund^b, Jeppe Lund Nielsen^b,

6 Per Halkjær Nielsen^{b,*}

7

8 ^a Department of Civil and Environmental Engineering, Graduate School of Engineering, Hiroshima

9 University, 1-4-1 Kagamiyama, Higashihiroshima, 739-8527 Japan.

10 ^b Department of Biotechnology, Chemistry and Environmental Engineering, Aalborg University,

11 Sohngaardsholmsvej 49, DK-9000 Aalborg, Denmark.

12

13 **Content:**

14

15 **Experimental Procedure**

16 **Figure S1**

17 **Table S1 – Table S3**

18

19

20

21

22

23

24

25

1 **Experimental Procedure**

2

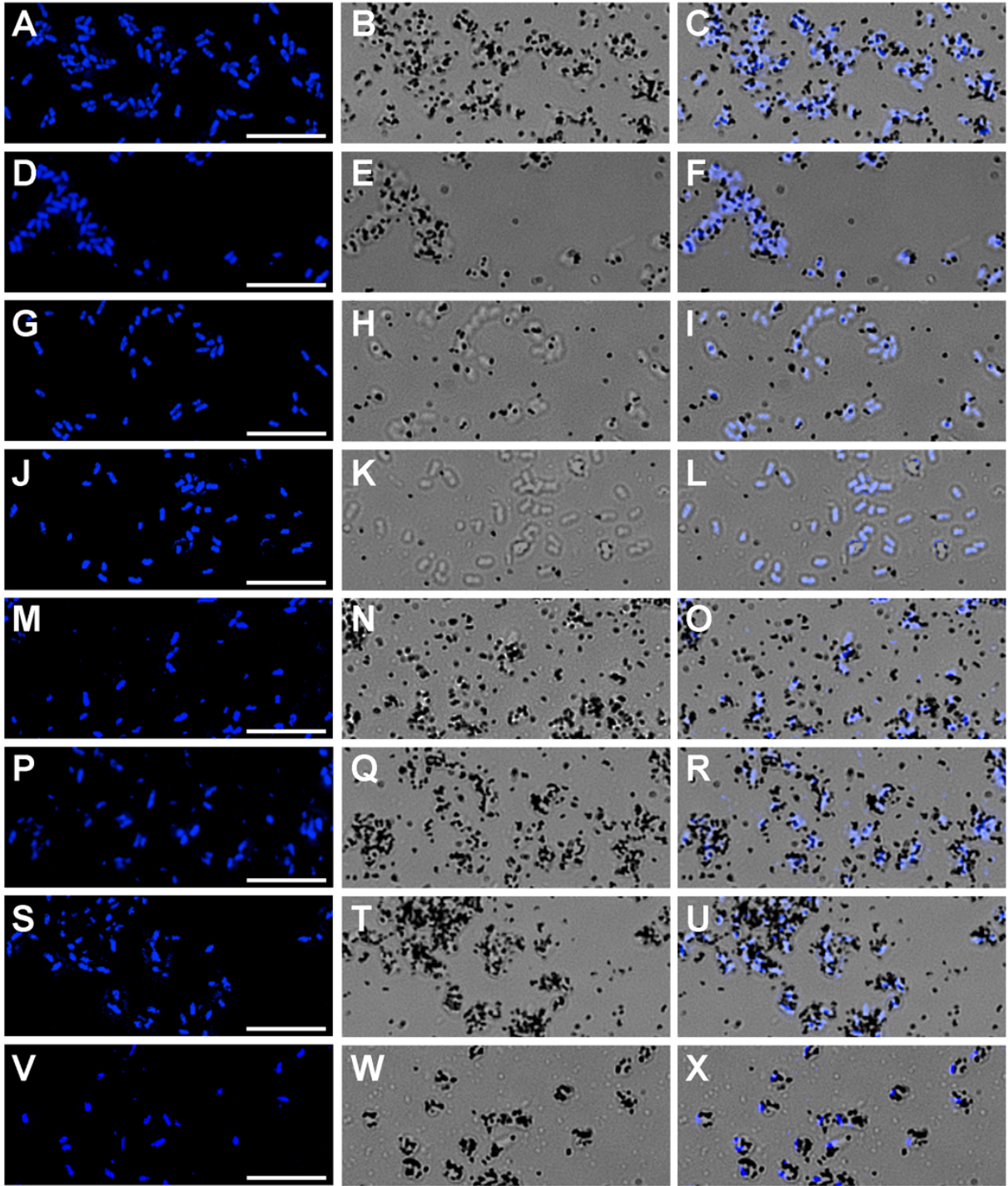
3 **Labeled organic substrates**

4 The labeled substrates under in this study were sodium [³H]acetate (NET003; specific activity: 8000 mCi
5 mmol⁻¹), sodium [2-¹⁴C]propionate (ARC0281A; specific activity: 55 mCi mmol⁻¹), sodium [2-
6 ¹⁴C]pyruvate (NEC256; specific activity: 16 mCi mmol⁻¹), [2-³H]glycerol (ART0188A; specific activity:
7 20000 mCi mmol⁻¹), [9,10(N)-³H]oleic acid (TRK140; specific activity: 7000 mCi mmol⁻¹), D-[6-
8 ³H(N)]glucose (NET100; specific activity: 32300 mCi mmol⁻¹), D-[2-³H(N)]mannose (NET570A;
9 specific activity: 21600 mCi mmol⁻¹), D-[1-³H]galactose (TRK233; specific activity: 7000 mCi mmol⁻¹),
10 [2-³H]glycine (TRK71; specific activity: 23000 mCi mmol⁻¹), L-[4,5-³H]leucine (TRK754; specific
11 activity: 64000 mCi mmol⁻¹), and [1-³H]ethanol (ART0335; specific activity: 5000 mCi mmol⁻¹). Labeled
12 substrates were purchased from PerkinElmer, Inc. (Waltham, MA), American Radiolabeled Chemicals,
13 Inc. (Saint Louis, MO), or Amersham Biosciences (Little Chalfont, United Kingdom).

14

1 **Figure S1**

2 Micrographs of *E. coli* MG1655 (stationary phase) incubated with [³H]glycerol under the multiple
3 substrates (acetate + glycerol) condition (A-C), with [³H]glycerol under the multiple substrates (oleic acid
4 + glycerol) condition (D-F), with [³H]glycerol under the high substrate concentration after starvation under
5 low O₂ concentration condition (G-I), with [³H]glycerol under the low substrate concentration after
6 starvation under low O₂ concentration condition (J-L), with [³H]glucose under the multiple substrates
7 (acetate + glucose) condition (M-O), with [³H]glucose under the multiple substrates (oleic acid + glucose)
8 condition (P-R), with [³H]glucose under the high substrate concentration after starvation under low O₂
9 concentration condition (S-U), and with [³H]glucose under the low substrate concentration after starvation
10 under low O₂ concentration condition (V-X). Panel A, D, G, J, M, P, S, and V show DAPI images. Panel
11 B, E, H, K, N, Q, T, and W are bright-field MAR images, corresponding to the DAPI images (A, D, G, J,
12 M, P, S, and V, respectively). Panel C, F, I, L, O, R, U, and X are overlay images showing strongly high
13 positive (C, O, R, U, and X), positive (F), some positive (I), and negative (L). The bars represent 10 μm.



2 Figure S1

1 **Table S1.** Substrate uptake patterns of the four probe-defined populations under aerobic conditions –
 2 previous studies from full-scale EBPR plants.

Substrate	Microthrix	Caldilinea	Accumulibacter	Azoarcus
Formate	-	-	-	-
Acetate	-	-	+	+
Propionate	-	-	+	(+)
Butyrate	-	-	-	-
Lactate	-	ND	ND	-
Pyruvate	-	-	+	+
Oleic acid	+	-	-	-
Glucose	-	+	-	-
Mannose	-	(+)	-	-
Galactose	-	(+)	-	-
N-acetyl-glucosamine	-	+	-	ND
Glycine	-	-	ND	-
Leucine	-	-	-/+*	-/+*
Glutamate	-	ND	+	-
Methanol	ND	ND	ND	+
Ethanol	-	-	-	(+)
Glycerol	ND	ND	ND	ND
References	Andreasen and Nielsen, 2000 Nielsen et al., 2002	Kragelund et al., 2011	Kong et al., 2004	Thomsen et al., 2007

4

5 ND: Not determined

6 *: uptake of leucine only in presence of acetate

7 () a minor fraction of probe-defined cells active

8

9

10 References

11 Andreasen, K., and Nielsen, P.H. (2000) Growth of *Microthrix parvicella* in nutrient removal activated sludge plants,
 12 studies of in situ physiology. *Water Res* **34**: 1559–1569.

13 Kragelund, C., Thomsen, T.R., Mielczarek, A.T., and Nielsen, P.H. (2011) Eikelboom's morphotype 0803 in
 14 activated sludge belongs to the genus *Caldilinea* in the phylum *Chloroflexi*. *FEMS Microbiol Ecol* **76**: 451–462.

15 Kong, Y., Nielsen, J.L., and Nielsen, P.H. (2004) Microautoradiographic study of *Rhodocyclus*-related poly-P
 16 accumulating bacteria in full-scale enhanced biological phosphorus removal plants. *Appl Environ Microbiol* **70**:
 17 5383–5390.

18 Nielsen, P.H., Roslev, P., Dueholm, T., and Nielsen, J.L. (2002) *Microthrix parvicella*, a specialized lipid consumer in
 19 anaerobic – aerobic activated sludge plants. *Water Sci Technol* **46**: 73–80.

20 Thomsen, T.R., Kong, Y., and Nielsen, P.H. (2007) Ecophysiology of dominant denitrifying bacteria in activated
 21 sludge. *FEMS Microbiol Ecol* **60**: 370–382.

22

1 **Table S2.** Experimental overview of high, low, and multiple substrate concentration conditions.
 2 Incubation time was for all 2 h. In all incubations the added labeled substrate was 5 $\mu\text{Ci ml}^{-1}$.

3

Experiment	Labeled substrate	Unlabeled substrate (s) concentration (mM)
High concentration	[³ H]acetate	2
	[¹⁴ C]propionate	2
	[¹⁴ C]pyruvate	2
	[³ H]oleic acid	0.5
	[³ H]glycerol	2
	[³ H]glucose	2
	[³ H]mannose	2
	[³ H]galactose	2
	[³ H]glycine	2
	[³ H]leucine	2
	[³ H]ethanol	0.5
Low concentration	[³ H]acetate	0.02
	[¹⁴ C]propionate	0.02
	[¹⁴ C]pyruvate	0.02
	[³ H]oleic acid	0.005
	[³ H]glycerol	0.02
	[³ H]glucose	0.02
	[³ H]mannose	0.02
	[³ H]galactose	0.02
	[³ H]glycine	0.02
	[³ H]leucine	0.02
	[³ H]ethanol	0.005
Multiple substrate condition	[¹⁴ C]propionate	Acetate (2 mM) + Propionate (2 mM)
	[¹⁴ C]pyruvate	Acetate (2 mM) + Pyruvate (2 mM)
	[³ H]oleic acid	Acetate (2 mM) + Oleic acid (0.5 mM)
	[³ H]glycerol	Acetate (2 mM) + Glycerol (2 mM)
	[³ H]glucose	Acetate (2 mM) + Glucose (2 mM)
	[³ H]mannose	Acetate (2 mM) + Mannose (2 mM)
	[³ H]galactose	Acetate (2 mM) + Galactose (2 mM)
	[³ H]glycine	Acetate (2 mM) + Glycine (2 mM)
	[³ H]leucine	Acetate (2 mM) + Leucine (2 mM)
	[³ H]ethanol	Acetate (2 mM) + Ethanol (0.5 mM)
	[³ H]acetate	Oleic acid (0.5 mM) + Acetate (2 mM)
	[¹⁴ C]propionate	Oleic acid (0.5 mM) + Propionate (2 mM)
	[¹⁴ C]pyruvate	Oleic acid (0.5 mM) + Pyruvate (2 mM)
	[³ H]glycerol	Oleic acid (0.5 mM) + Glycerol (2 mM)
	[³ H]glucose	Oleic acid (0.5 mM) + Glucose (2 mM)
	[³ H]mannose	Oleic acid (0.5 mM) + Mannose (2 mM)
	[³ H]galactose	Oleic acid (0.5 mM) + Galactose (2 mM)
	[³ H]glycine	Oleic acid (0.5 mM) + Glycine (2 mM)
	[³ H]leucine	Oleic acid (0.5 mM) + Leucine (2 mM)
	[³ H]ethanol	Oleic acid (0.5 mM) + Ethanol (0.5 mM)
	[³ H]acetate	Glucose (2 mM) + Acetate (2 mM)
	[¹⁴ C]propionate	Glucose (2 mM) + Propionate (2 mM)
	[¹⁴ C]pyruvate	Glucose (2 mM) + Pyruvate (2 mM)
	[³ H]oleic acid	Glucose (2 mM) + Oleic acid (0.5 mM)
	[³ H]glycerol	Glucose (2 mM) + Glycerol (2 mM)
	[³ H]mannose	Glucose (2 mM) + Mannose (2 mM)
	[³ H]galactose	Glucose (2 mM) + Galactose (2 mM)
	[³ H]glycine	Glucose (2 mM) + Glycine (2 mM)
	[³ H]leucine	Glucose (2 mM) + Leucine (2 mM)
	[³ H]ethanol	Glucose (2 mM) + Ethanol (0.5 mM)

4

1 **Table S3.** Experimental overview of starvation and induction conditions.

Experiment	Unlabeled substrate	Incubation time (h)	O ₂ concentration in headspace (%)	Incubation with labeled substrate, ^a high (0.5-2 mM) or low (0.005-0.02 mM) concentration
Starvation under high O ₂	None	12	20	All substrates, high and low concentration
Starvation under low O ₂	None	12	1	All substrates, high and low concentration
Induction with glucose	Glucose (3 mM)	12	20	Glucose, mannose, or galactose, high concentration
Induction with mannose	Mannose (3 mM)	12	20	Glucose, mannose, or galactose, high concentration
Induction with galactose	Galactose (3mM)	12	20	Glucose, mannose, or galactose, high concentration
Induction with leucine	Leucine (3 mM)	12	20	Glycine, leucine, high concentration

2 ^a After starvation and induction, samples were incubated with labeled substrate as shown in Table S2.

3