

Enrichment Using an Up-flow Column Reactor and Community Structure of Marine Anammox Bacteria from Coastal Sediment

TOMONORI KINDAICHI^{1*}, TAKANORI AWATA¹, YUJI SUZUKI², KATSUICHIRO TANABE³, MASASHI HATAMOTO⁴, NORIATSU OZAKI¹, and AKIYOSHI OHASHI¹

¹Department of Civil and Environmental Engineering, Graduate School of Engineering, Hiroshima University, 1-4-1 Kagamiyama, Higashihiroshima 739-8527, Japan; ²Wakayama Prefectural Bureau, 1-1 Komatsubaradori, Wakayama 640-8585, Japan; ³Matsuo Corporation, 1-10 Hachimankoji, Saga 840-8666, Japan; and ⁴Department of Environmental Systems Engineering, Nagaoka University of Technology, 1603-1 Kamitomioka, Nagaoka 940-2188, Japan

(Received August 18, 2010—Accepted November 8, 2010—Published online December 11, 2010)

We established an enrichment culture of marine anaerobic ammonium oxidation (anammox) bacteria using an up-flow column reactor fed with artificial sea water supplemented with nitrogen and minerals and inoculated with coastal surface sediment collected from Hiroshima Bay. After 2 months of reactor operation, simultaneous removal of NH_4^+ and NO_2^- was observed, suggesting that an anammox reaction was proceeding. A total nitrogen removal rate of $2.17 \text{ g-N L}^{-1} \text{ day}^{-1}$ was attained on day 594 while the nitrogen loading rate was $3.33 \text{ g-N L}^{-1} \text{ day}^{-1}$. Phylogenetic analysis revealed that at least two dominant “*Candidatus Scalindua*” species were present in this reactor. Moreover, many uncultured bacteria and archaea, including candidate division or ammonia-oxidizing archaea, were present. Fluorescence *in situ* hybridization (FISH) revealed that anammox bacteria accounted for $85.5 \pm 4.5\%$ of the total bacteria at day 393. We also designed two oligonucleotide probes specific to each dominant “*Candidatus Scalindua*” species. A simultaneous FISH analysis using both probes showed that two different “*Candidatus Scalindua*” species were clearly recognizable and coexisted during reactor operation, although there was some variation in their abundance. The marine anammox bacteria enriched in this study have potential applications to the treatment of industrial wastewater containing high levels of ammonium and salt.

Key words: anaerobic ammonium oxidation (anammox), marine anammox bacteria, “*Candidatus Scalindua*” species, enrichment culture, FISH, phylogenetic analysis

Anaerobic ammonium oxidation (anammox) is a biological process in which ammonium is directly converted to dinitrogen gas with nitrite as the electron acceptor under anoxic conditions (9). Anammox is mediated by a group of deep-branching *Planctomycetes*-like bacteria (31). The anammox process is an efficient and cost-effective alternative to conventional processes used to remove nitrogen from ammonia-rich wastewater (9). The advantages of the anammox process over the conventional combination of nitrification and denitrification processes include lower oxygen demand and no requirement for external carbon sources. Because some types of industrial wastewater, such as landfill leachate, contain high salt concentrations, salinity is an important parameter for wastewater treatment (11). Anammox bacteria have been detected in wastewater treatment facilities and natural environments worldwide (26). Currently, at least five genera of anammox bacteria have been reported and proposed, including “*Candidatus Brocadia*”, “*Candidatus Kuenenia*”, “*Candidatus Scalindua*”, “*Candidatus Anammoxoglobus*”, and “*Candidatus Jettenia*”.

The “*Candidatus Scalindua*” group is primarily found in marine environments such as the Black Sea (16) and the Arabian Sea, as well as along the coasts of Namibia, Chile, Peru and Japan (3, 37). It is currently estimated that anammox bacteria are responsible for at least 30–50% of nitrogen

removal from marine ecosystems (6), although this value is not undisputed (36). Since neither marine anammox bacteria nor wastewater anammox bacteria have been isolated in pure cultures, enrichment of marine anammox bacteria is valuable for understanding a microbial community in association with the marine anammox reaction and *in situ* activities of marine anammox bacteria and for application to the treatment of industrial wastewater containing high levels of ammonium and salt. Although “*Candidatus Scalindua*” is dominant in marine environments, only a few “*Candidatus Scalindua*” species have been enriched from sea sediment (12, 20, 35). Furthermore, the cultures showed a low population density (67–90% of total bacteria) and/or low conversion rates ($0.07\text{--}0.35 \text{ g L}^{-1} \text{ day}^{-1}$) after long periods of enrichment.

We previously reported that anammox bacteria derived from wastewater were successfully enriched using up-flow fixed-bed column reactors, and that the anammox reaction was observed within 2–3 months after the reactor start-up (14, 33). These studies were conducted based on strategies specific for the selected inoculum and a short hydraulic retention time (HRT). The short HRT might prevent substrate limitation and the accumulation of inhibitory substances, as well as establish enrichment with low concentrations of media.

The present study was conducted to establish an enrichment culture of anammox bacteria in a marine environment with a high conversion rate and to investigate whether enriched anammox bacteria belong to the “*Candidatus*

* Corresponding author. E-mail: tomokin@hiroshima-u.ac.jp;
Tel: +81-82-424-5718; Fax: +81-82-424-5718.

Scalindua" group. In addition, the types of bacteria and archaea that coexisted in the enriched culture were investigated. To achieve these objectives, an up-flow fixed-bed column reactor with a short HRT was used with coastal sediment collected from Hiroshima Bay as the inoculum. Phylogenetic analysis based on 16S rRNA genes and fluorescence *in situ* hybridization (FISH) were performed to investigate the community structure of enriched anammox bacteria together with coexisting microorganisms and to monitor long-term changes of dominant anammox bacterial populations.

Materials and Methods

Sediment sample

A surface sediment sample was collected from Hiroshima Bay (34°21.4'N, 132°30.7'E, water depth 7.4 m) in the Seto Inland Sea during July 2008 using a plastic core sampler. The bottom water temperature, pH, salinity, ammonia, nitrite, and nitrate concentrations were 19.4°C, 8.0, 32.1 psu, 17.9 µM, 2.9 µM, and 6.4 µM, respectively.

Enrichment

A surface sediment sample weighing 1 g (wet weight) was used as an inoculum in a glass column reactor (diameter, 2.6 cm; length, 10.5 cm) with a nonwoven fabric sheet (Japan Vilene, Tokyo, Japan) as the biofilm carrier material. To confirm the existence of DNA derived from anammox bacteria in the sediment sample, PCR was conducted using *Planctomycetales*-specific primer sets and the DNA was electrophoresed prior to inoculation (see section *DNA extraction and PCR amplification*). The reactor volume was 56 cm³, the surface area of the biofilm carrier was 54.6 cm², the temperature was maintained at 20°C, and the initial hydraulic retention time (HRT) was 0.9 h. A synthetic marine nutrient medium composed of the following was used: 35 g L⁻¹ of SEALIFE (Marine Tech, Tokyo, Japan), an artificial sea salt used in this study, supplemented with (NH₄)₂SO₄ (24 to 88 mg L⁻¹), NaNO₂ (25 to 88 mg L⁻¹), KHCO₃ (500 mg L⁻¹), KH₂PO₄ (27 mg L⁻¹), MgSO₄·7H₂O (300 mg L⁻¹), CaCl₂·2H₂O (180 mg L⁻¹), and 1 mL of trace element solutions I and II (34). It should be noted that 35 g L⁻¹ of SEALIFE contained the following: Na (9.14 g L⁻¹), Mg (1.23 g L⁻¹), Ca (0.34 g L⁻¹), K (0.35 g L⁻¹), Cl (16.89 g L⁻¹), SO₄²⁻ (2.3 g L⁻¹), Sr (0.007 g L⁻¹), Br (0.055 g L⁻¹), F (0.2 mg L⁻¹), B (0.9 mg L⁻¹), Li (0.16 mg L⁻¹), V (0.003 mg L⁻¹), Rb (0.11 mg L⁻¹), Mo (0.012 mg L⁻¹), and Ba (0.016 mg L⁻¹). The medium was flushed with N₂ gas for at least 1 h before the nutrients were added to achieve a concentration of dissolved oxygen (DO) below 0.5 mg L⁻¹. The total nitrogen loading and removal rates were calculated based on the concentrations of NH₄⁺, NO₂⁻, and NO₃⁻ as well as the HRT.

Analytical methods

The seawater quality (see above) was determined using a multi-parameter water quality meter (AAQ1183, JFE Alec, Kobe, Japan). The NH₄⁺ concentration was determined using Nessler's method with a UV-visible spectrophotometer (DR-2800, Hach, Loveland, CO, USA). The concentrations of NO₂⁻ and NO₃⁻ were determined using ion-exchange chromatography (HPLC 10Avp; Shimadzu, Kyoto, Japan) with a Shodex Asahipak NH2P-50 4D anion column (Showa Denko, Tokyo, Japan) and UV-VIS detector (SPD-10A, Shimadzu) after filtration of samples through 0.2-µm pore-size membranes (Advantec, Tokyo, Japan).

DNA extraction and PCR amplification

Total DNA was extracted from the anammox biofilm in the column reactor after 222 days of operation using a Fast DNA spin kit (MP Biomedicals, Irvine, CA, USA) according to the manufacturer's instructions. To construct three clone libraries, an anammox

bacterial clone library, a total bacterial clone library, and an archaeal clone library, the 16S rRNA gene fragments from the isolated total DNA were amplified using a ONE Shot LA PCR MIX kit (Takara Bio, Ohtsu, Japan) and a *Planctomycetales*-specific primer set, pla46f (21) and univ1390r (38), a bacteria-specific primer set, eub338f (an equimolar mixture of complementary sequences of the EUB338, EUB338 I, EUB338 II, and EUB338 III probes (2, 5)) and univ1390r (38), or an archaea-specific primer set, arc109f and arc912r (18) (Table S1). The PCR conditions for the anammox bacteria were as follows: 4 min of initial denaturation at 94°C, followed by 30 cycles of 45 s at 94°C, 50 s at 58°C, and 3 min at 72°C. The final extension was conducted for 10 min at 72°C. The PCR conditions for the total bacteria were as follows: 5 min of initial denaturation at 94°C, followed by 25 cycles of 30 s at 94°C, 30 s at 50°C, and 1.5 min at 72°C. The final extension was conducted for 5 min at 72°C. The PCR conditions for the archaea were as follows: 5 min of initial denaturation at 94°C, followed by 25 cycles of 30 s at 94°C, 30 s at 50°C, and 1 min at 72°C. The final extension was conducted for 5 min at 72°C. All PCRs were performed using a total volume of 50 µL containing 1 µg of DNA as the template. The PCR products were electrophoresed in a 1% (w/v) agarose gel.

Cloning and phylogenetic analysis

The PCR products were purified using a QIAquick PCR Purification Kit (Qiagen, Hilden, Germany) and then cloned using a TOPO XL PCR Cloning Kit (Invitrogen, Carlsbad, CA, USA) according to the manufacturer's instructions. Cloned 16S rRNA genes were randomly selected from each clone library and subjected to sequencing. DNA sequencing was performed by Dragon Genomics Center (Takara Bio, Yokkaichi, Japan). The sequences were compared with sequences of the reference organisms using a BLAST search (1). Sequences with 97% or higher similarity were grouped into operational taxonomic units (OTUs) using the Similarity Matrix program from the Ribosomal Database Project (19). The 16S rRNA gene sequences from the anammox bacterial clone library (1,385 bp), total bacterial clone library (1,070 bp), and archaeal clone library (788 bp) were imported and aligned using Integrated Aligners in the ARB software (17), and alignments were refined manually. Phylogenetic trees were constructed using neighbor-joining (24) with Olsen correction model and maximum parsimony (Phylip DNAPARS) approaches using default settings in the ARB software with the greengenes236469 database (<http://greengenes.lbl.gov/cgi-bin/nph-index.cgi>). A bootstrap resampling analysis (7) of 1,000 replicates was performed using the ARB software to estimate the confidence of tree topologies.

Oligonucleotide probes, FISH, and cell counting

The 16S rRNA-targeted oligonucleotide probes used in this study were EUB338 (2), EUB338 II (5), and EUB338 III (5) for total bacteria, Amx368 (25), S^{*}-Sca-1129-a-A-18 (5'-TACCCGG CACAACCCGCT-3'), and S^{*}-Sca-1129-b-A-18 (5'-TACTCGG CATTACCCGAT-3') for anammox bacteria, and ARC915 (29), Cren512 (10), and EURY514 (10) for archaea (S-Table 1). Probes S^{*}-Sca-1129-a-A-18 and S^{*}-Sca-1129-b-A-18 were designed in this study using the PROBE_DESIGN tool of the ARB software (17) with the greengenes236469 database. The specificity of the probes was checked based on a comparative analysis of all sequences in the ARB database which comprised both publicly available sequences and our clone sequences. To determine the formamide concentration for optimum stringency, the designed probes were analyzed on a paraformaldehyde-fixed anammox biofilm applying hybridization buffer containing 0–60% formamide (5% increments). The probes were labeled with fluorescein isothiocyanate (FITC) or tetramethylrhodamine 5-isothiocyanate (TRITC). Dehydration and FISH were performed as described previously (23). Hybridization with the probes requiring different stringency conditions was conducted using successive procedures; hybridization with the probe requiring higher stringency was always

performed first. Either a LSM5 PASCAL confocal laser-scanning microscope equipped with an Ar ion laser (488 nm) and an HeNe laser (543 nm) (Carl Zeiss, Oberkochen, Germany), or an Axioimager M1 epifluorescence microscope (Carl Zeiss) was used for observation of the samples.

For the quantitative determination of microbial composition in the anammox biofilm, surface fractions of the specific probe-hybridized cell area and EUB338mix (composed of EUB338, EUB338 II, and EUB338 III) probe-hybridized cell area were determined following homogenization of biofilm samples and simultaneous *in situ* hybridizations with various probe sets. The average surface fraction was determined from at least 10 representative laser scanning micrographs obtained from each biofilm sample using image analysis software provided by Zeiss (13).

The total number of cells was determined by 4',6'-diamidino-2-phenylindole (DAPI) direct counting. Briefly, marine anammox biofilm samples (approximately 1.0 cm²) were collected from the reactor, homogenized, and suspended in 1 mL of phosphate-buffered saline (PBS) (10 mM sodium phosphate buffer and 130 mM sodium chloride; pH 7.2).

Nucleotide sequence accession numbers

The GenBank/EMBL/DBJ accession numbers for the 16S rRNA gene sequences of the 47 OTU sequences used for the phylogenetic tree are AB573102 to AB573148.

Results and Discussion

Enrichment of marine anammox bacteria

An up-flow fixed-bed column reactor was used for the enrichment of marine anammox bacteria in this study. Before the enrichment, we examined the existence of DNA derived from anammox bacteria in the sediment sample to select an appropriate inoculum for rapid start-up of the anammox

reactor. The DNA band amplified by the *Planctomycetales*-specific primers was observed by electrophoresis in the 1% agarose gel (data not shown). Consequently, one gram of the sediment was inoculated on a nonwoven fabric sheet in the glass column reactor. Upon start-up of the enrichment, the synthetic marine nutrient medium containing 5 mg-N L⁻¹ of ammonium and 5 mg-N L⁻¹ of nitrite was fed to the column reactor with an HRT of 0.9 h. This feeding condition was maintained for 4 months. The simultaneous removal of NH₄⁺ and NO₂⁻ was observed after 68 days of incubation (Fig. 1A), suggesting that an anammox reaction was occurring in the reactor. During this period, the NH₄⁺ and NO₂⁻ removal efficiencies were 30% and 59%, respectively. A 97% NO₂⁻ removal efficiency was observed on day 114 (Fig. 1A). After day 114, the nitrogen loading rate gradually increased as the NH₄⁺ and NO₂⁻ concentrations increased (Fig. 1B). It should also be noted that higher nitrogen loading rates on days 172 and 276 were due to pump failure. When the nitrogen removal rate dropped suddenly, the nitrogen loading rate decreased due to a decreasing flow rate (increasing HRT); however, the nitrogen removal rate then gradually recovered. The color of the biofilm changed from gray to khaki after about 60 days and from khaki to orange after 300 days. After 300 days of reactor operation, the nitrogen loading rate gradually increased as the HRT was reduced to 0.3 h (Fig. 1B). A total nitrogen removal rate of 2.17 g-N L⁻¹ day⁻¹ was attained on day 594, while the nitrogen loading rate was 3.33 g-N L⁻¹ day⁻¹ (HRT, 0.3 h). The average nitrogen stoichiometric ratio after 300 days of operation was 1:1.28:0.24 for consumption of NH₄⁺ and NO₂⁻ for the production of NO₃⁻. This stoichiometric ratio was similar to that

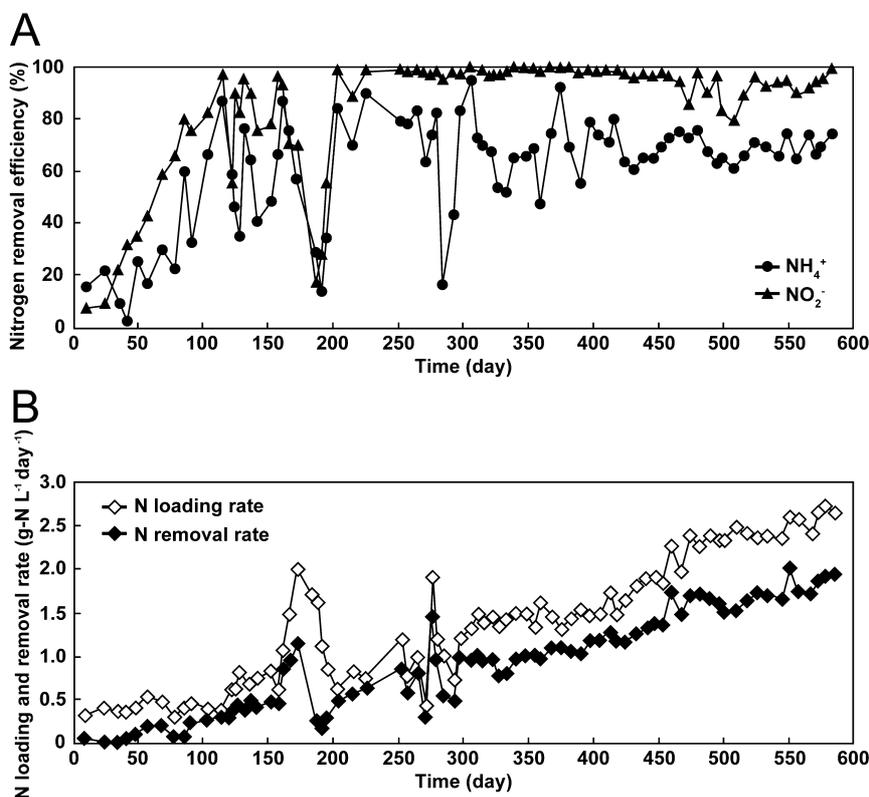


Fig. 1. NH₄⁺ and NO₂⁻ removal efficiency (A) and nitrogen loading rate and removal rate (B) during the experiment.

reported previously, 1:1.32:0.26, for the anammox reaction in a sequencing batch reactor fed synthetic wastewater (30).

The nitrogen removal rate was six to 31 times higher than the values reported in other studies that used a column reactor fed deep sea water containing minerals (20), a column reactor fed a synthetic salt-based medium (12), and sequencing batch reactors fed a sea salt-based medium (35). The high nitrogen removal rate in this study was attributed to the short HRT. Specifically, the 0.3 to 0.9 h HRT employed was much shorter than in other studies (12, 20, 35). We previously reported that a reduction of HRT might prevent nitrite limitation and the accumulation of unknown by-products derived from the anammox reaction (33). Other studies have also found that the performance of anammox reactors was enhanced by stepwise reductions of the HRT (27).

We successfully enriched marine “*Candidatus Scalindua*” species involved in marine anammox using a synthetic marine nutrient medium containing 35 g L⁻¹ of artificial sea salt (SEALIFE). Van de Vossenberg *et al.* (35) reported that two species of marine anammox bacteria were successfully enriched when a Red Sea salt-based medium was used. Nakajima *et al.* (20) also established an enrichment culture of marine anammox bacteria using deep sea water. Conversely, the addition of salt (NaCl and KCl) to standard anammox medium (34) did not result in the enrichment of marine anammox bacteria (11). Due to the low growth rate of marine anammox bacteria, we did not attempt to enrich the bacteria using medium that contained individual salts. However, the presence of microelements that are normally contained in sea water such as bromine, strontium, and fluorine may enhance the enrichment. In fact, Kawagoshi *et al.* (12) reported the successful enrichment of “*Candidatus Scalindua*” species using synthetic sea water.

Microbial community structure of the marine anammox bacterial biofilm

To reveal the structure of microbial communities residing in the glass column reactor, three 16S rRNA gene clone libraries, an anammox bacterial clone library, a total bacterial clone library, and an archaeal clone library, were constructed from the biofilm samples taken from the reactor on day 222 at which time a stable anammox reaction after the first pump failure was observed. No chimeric sequences were detected in any of the clone libraries. We obtained 93, 90, and 88 clones from the anammox bacterial clone library, the total bacterial clone library, and archaeal clone library, respectively. The clones were grouped into OTUs based on 97% or higher sequence similarity. We identified at least 12 bacterial phyla and two archaeal phyla including 47 different operational taxonomic units (OTUs) in the marine anammox biofilm. It should be noted that there were no differences in the topology of the two trees generated using the neighbor-joining and the maximum parsimony methods.

In the anammox bacterial phylogenetic tree, 93 clones were classified into two groups (Fig. 2). The sequence of clone husup-a2 was closely related to that of an uncultured *Planctomycetales* bacterium, clone Peru_45, with 98% similarity. This clone was also related to a proposed anammox bacterium, “*Candidatus Scalindua brodae*”, with 97% sequence similarity. The sequence similarity of the clone

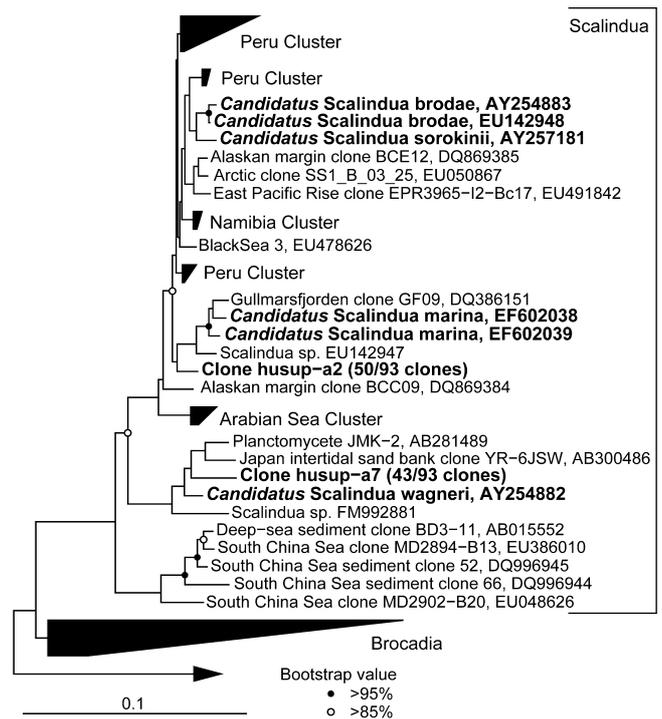


Fig. 2. Phylogenetic tree of clones obtained from the enrichment and related anammox bacteria. The tree was generated by the neighbor-joining method. The scale bar represents the number of nucleotide changes per sequence position. The symbols at each branch point show the bootstrap values obtained with 1,000 resamplings. The GenBank/EMBL/DBJ accession numbers are also indicated. *Candidatus* genera designations are bracketed on the right. The *Planctomyces maris* (AJ231184) sequence served as the outgroup for rooting the tree. The numbers in parentheses indicate the frequencies of appearance of identical clones among the clones analyzed.

husup-a7 to the closest proposed anammox bacterium, “*Candidatus Scalindua wagneri*”, was less than 96%. According to Stackebrandt and Goebel (28), the dominant anammox-like bacterium detected in this study was most likely a novel species of the genus “*Candidatus Scalindua*”. The sequence similarity between husup-a2 and husup-a7 was only 93.1%. A distance matrix for the two dominant anammox species indicated that clone husup-a2 belonged to “*Candidatus Scalindua 1*” (“*Candidatus Scalindua sorokinii/brodae*”, (37)), whereas clone husup-a7 was closely related to “*Candidatus Scalindua wagneri*”. These results clearly indicate that at least two different “*Candidatus Scalindua*” species were present in this reactor.

Ninety clones in the total bacterial phylogenetic tree were classified into 42 OTUs (12 phyla). Figure 3 shows the phylogenetic tree of bacteria without the *Planctomycetes* group. The percentage of bacterial clones other than *Planctomycetes* exceeded 80% of all the clones detected in this study. In contrast to the anammox bacterial clone library, the total bacterial clone library showed a high diversity of clones at the phylum level. These clones were primarily related to uncultured bacterial clones detected in surface sediment from hydrothermal vents, the Arctic Ocean, or hypersaline mats with 87 to 99% sequence similarity; thus, the physiological characteristics of these uncultured bacteria are not clear.

In the archaeal phylogenetic tree, 88 clones were classified into nine groups (Fig. 4). The clones related to

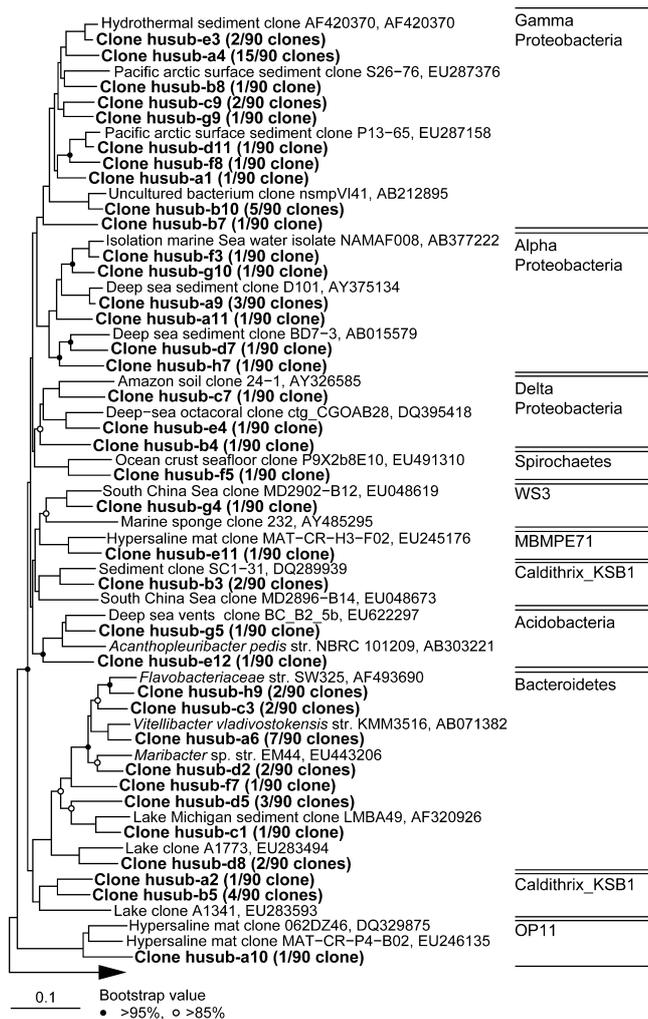


Fig. 3. Phylogenetic tree of clones obtained from the enrichment and related bacteria. The tree was generated by the neighbor-joining method. The scale bar represents the number of nucleotide changes per sequence position. The symbols at each branch point show the bootstrap values obtained with 1,000 resamplings. The GenBank/EMBL/DBJ accession numbers are also indicated. Division and subdivision designations are bracketed on the right. The *Aquifex aeolicus* (AJ309733) sequence served as the outgroup for rooting the tree. The numbers in parentheses indicate the frequencies of appearance of identical clones among the clones analyzed.

the *Euryarchaeota* and *Crenarchaeota* accounted for 98% and 2% of the total, respectively. All clones except for husua-c3 were affiliated with uncultured archaea with 84 to 94% sequence similarity and belonged to a cluster of clones belonging to deep sea hydrothermal vent euryarchaeotal group 6 (DHVEG-6). This group has primarily been detected in ocean water, coastal water, polar seawater, and deep-sea hydrothermal vents (22, 32). Unfortunately, the ecophysiology and functions of the coexisting *Euryarchaeota* are not yet clear. The sequences of clone husua-c3 and *Nitrosopumilus maritimus* exhibited 99% similarity. *N. maritimus* was recently reported to grow chemolithoautotrophically through the aerobic oxidation of ammonia to nitrite, and is recognized as a member of the ammonia-oxidizing archaea (AOA) (15). AOA have also been found in water and sand samples collected from other bay areas (4, 8).

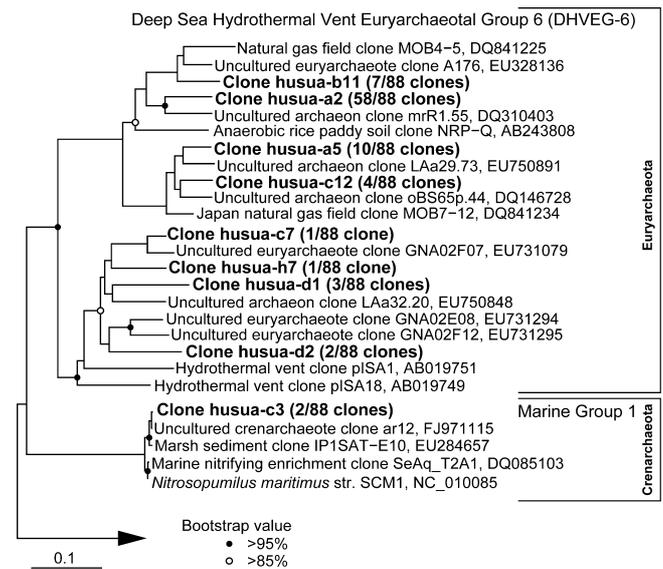


Fig. 4. Phylogenetic tree of clones obtained from the enrichment and related archaea. The tree was generated by the neighbor-joining method. The scale bar represents the number of nucleotide changes per sequence position. The symbols at each branch point show the bootstrap values obtained with 1,000 resamplings. The GenBank/EMBL/DBJ accession numbers are also indicated. Subdivision designations are bracketed on the right. The *Aquifex aeolicus* (AJ309733) sequence served as the outgroup for rooting the tree. The numbers in parentheses indicate the frequencies of appearance of identical clones among the clones analyzed.

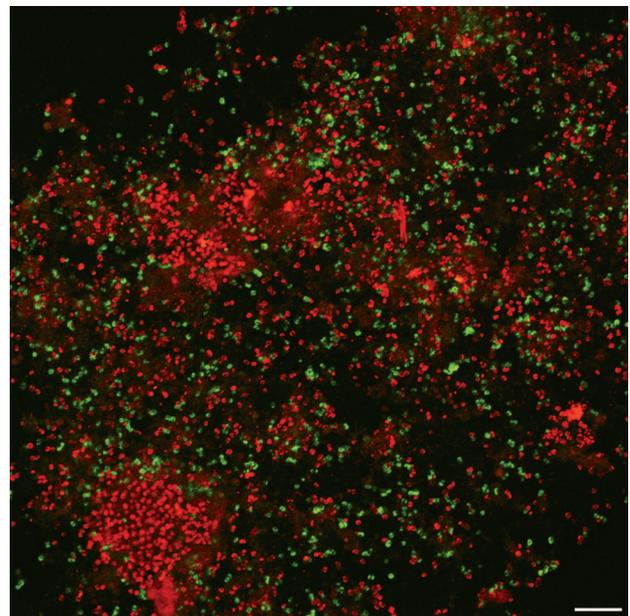


Fig. 5. Confocal laser scanning micrograph of the marine anammox biofilm after homogenization. Green corresponds to the FISH signal of the FITC-labeled probe S-*.Sca-1129-a-A-18. Red corresponds to the FISH signal of the TRITC-labeled probe S-*.Sca-1129-b-A-18. The bar represents 10 μm.

In situ detection and succession of marine anammox bacteria

First, we attempted to observe two different “*Candidatus Scalindua*” species by FISH using two newly designed probes, S-*.Sca-1129-a-A-18 and S-*.Sca-1129-b-A-18, which were specific for clones husup-a2 and husup-a7,

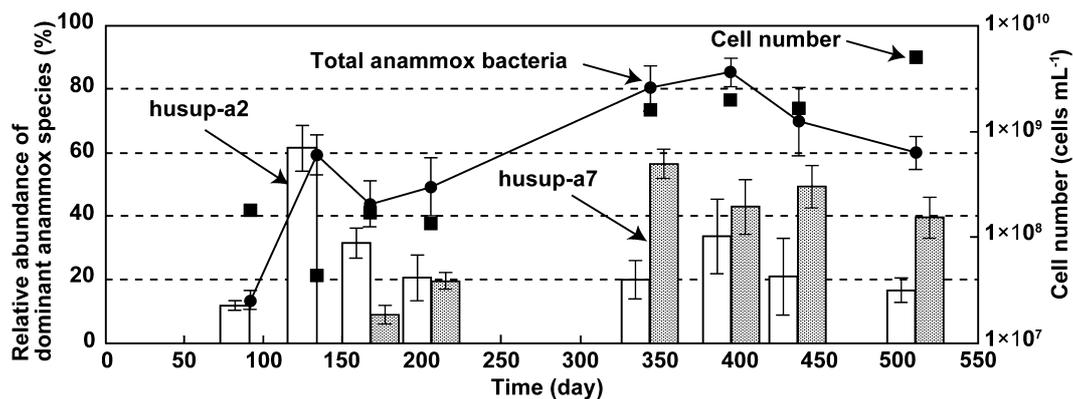


Fig. 6. Relative abundance of the dominant species husup-a2 (white bars) and husup-a7 (gray bars), and all anammox bacteria detected with the probe Amx368 (closed circles). Cell numbers (closed squares) were determined by DAPI counting. Relative abundance is shown as the percentage of each specific probe signal in a microscopic field to the EUB338mix probe signal. The error bars indicate standard deviations.

respectively. Then, changes in the populations of dominant anammox bacteria were determined by FISH. The optimal formamide concentrations were experimentally determined to be 35% for probe S^{*}-Sca-1129-a-A-18 and 20% for probe S^{*}-Sca-1129-b-A-18. Figure 5 shows a FISH image of the homogenized biofilm after 433 days of operation. *In situ* hybridization revealed that two different “*Candidatus Scalindua*” species were clearly recognized by the simultaneous use of probes S^{*}-Sca-1129-a-A-18 and S^{*}-Sca-1129-b-A-18. The two anammox bacteria enriched in this study showed a typical morphology (approximately 1 μm in diameter and doughnut-shaped) as reported by Kuypers *et al.* (16). No archaea were detected by FISH with the probe ARC915, Cren512, or EURY514, even though many groups of archaea were detected by the phylogenetic analysis. This was likely due to the archaeal population being present at levels below the detection limit of FISH and/or low activity in the biofilm. Another explanation involves the problem with penetration of probes into the cells as FISH was applied without any enzymatic treatments enhancing cell-wall permeability.

Changes in the populations of marine anammox bacteria were determined by FISH using the two newly designed probes (Fig. 6). The population of total anammox bacteria (detected with probe Amx368) as well as total number of cells increased with increasing operation time. Overall, the anammox bacteria accounted for $85.5 \pm 4.5\%$ of all the bacteria at day 393. Clone husup-a2 was found to be present throughout the period of operation and to account for 12.7 ± 1.3 to $62.2 \pm 7.3\%$ of the total. Clone husup-a7 was not observed until day 166, and comprised 9.8 ± 2.8 to $56.7 \pm 4.4\%$ of bacteria. The relative abundance of husup-a2 peaked at day 132, whereas that of husup-a7 peaked at day 342. The composition of total anammox bacteria corresponded with the sum of the compositions of clones husup-a2 and -a7. After day 166, two most abundant species of anammox bacteria coexisted during reactor operation (Fig. 6). One possible explanation for the presence of two species is the operational temperature, which was 20°C, corresponding to the temperature at the location from which the surface sediment sample was collected. Van de Vossenberg *et al.* (35) reported that two “*Candidatus Scalindua*” species were equally abundant in a 15°C culture, whereas one species of

“*Candidatus Scalindua*” gradually disappeared in a 23°C culture, while only one species of “*Candidatus Scalindua*” was dominant in a 30°C culture (12). Accordingly, the 20°C reactor temperature employed in this study might not be a selective condition for the two enriched species of “*Candidatus Scalindua*”. Thus, the temperature condition is likely to have a strong influence on the appearance of dominant species within “*Candidatus Scalindua*”. Accordingly, further study of the population dynamics of the two “*Candidatus Scalindua*” species at different temperatures should be performed.

In conclusion, we successfully enriched marine anammox species using a glass column reactor inoculated with coastal sediment collected from Hiroshima Bay and achieved a total nitrogen removal rate of $2.17 \text{ g-N L}^{-1} \text{ day}^{-1}$. FISH and phylogenetic analyses revealed that two “*Candidatus Scalindua*” species coexisted in the biofilm and accounted for about 80% of the total bacterial population at days 342 and 393. In addition, a high diversity of uncultured bacteria and archaea coexisted in the marine anammox biofilm. More studies are needed to investigate the characteristics of these marine anammox bacteria, such as the optimal growth temperature, pH, and salinity.

Acknowledgements

This research was partially supported by an Assistance for Research and Development of Construction Technology grant from the Ministry of Land, Infrastructure, Transport and Tourism, by a Grant-in-Aid for Scientific Research (C) from the Ministry of Education, Culture, Sports, Science and Technology, and by research foundations from the Institute for Fermentation (IFO), Osaka and Public Association for Construction Service Chugoku region. This study was conducted at the Analysis Center of Life Science, Hiroshima University. We thank the staff at the Coastal Engineering Laboratory, Hiroshima University, for providing coastal sediment.

References

- Altschul, S.F., W. Gish, W. Miller, E.W. Myers, and D.J. Lipman. 1990. Basic local alignment search tool. *J. Mol. Biol.* 215:403–410.
- Amann, R.L., L. Krumholz, and D.A. Stahl. 1990. Fluorescent-oligonucleotide probing of whole cells for determinative, phylogenetic, and environmental-studies in microbiology. *J. Bacteriol.* 172:762–770.

3. Amano, T., I. Yoshinaga, K. Okada, T. Yamagishi, S. Ueda, A. Obuchi, Y. Sako, and Y. Suwa. 2007. Detection of anammox activity and diversity of anammox bacteria-related 16S rRNA genes in coastal marine sediment in Japan. *Microbes Environ.* 22:232–242.
4. Ando, Y., T. Nakagawa, R. Takahashi, K. Yoshihara, and T. Tokuyama. 2009. Seasonal changes in abundance of ammonia-oxidizing archaea and ammonia-oxidizing bacteria and their nitrification in sand of an eelgrass zone. *Microbes Environ.* 24:21–27.
5. Daims, H., A. Bruhl, R. Amann, K.H. Schleifer, and M. Wagner. 1999. The domain-specific probe EUB338 is insufficient for the detection of all Bacteria: Development and evaluation of a more comprehensive probe set. *Syst. Appl. Microbiol.* 22:434–444.
6. Devol, A.H. 2003. Nitrogen cycle: Solution to a marine mystery. *Nature* 422:575–576.
7. Felsenstein, J. 1985. Confidence limits on phylogenies: An approach using the bootstrap. *Evolution* 39:783–791.
8. Hao, D.M., T. Tashiro, M. Kato, R. Sohrin, T. Ishibashi, C. Katsuyama, K. Nagaosa, H. Kimura, T.D. Thanh, and K. Kato. 2010. Population dynamics of *Crenarchaeota* and *Euryarchaeota* in the mixing front of river and marine waters. *Microbes Environ.* 25:126–132.
9. Jetten, M.S.M., M. Strous, K.T. van de Pas-Schoonen, *et al.* 1998. The anaerobic oxidation of ammonium. *FEMS Microbiol. Rev.* 22:421–437.
10. Jurgens, G., F.O. Glöckner, R. Amann, A. Saano, L. Montonen, M. Likolammi, and U. Münster. 2000. Identification of novel Archaea in bacterioplankton of a boreal forest lake by phylogenetic analysis and fluorescent *in situ* hybridization. *FEMS Microbiol. Ecol.* 34:45–56.
11. Kartal, B., M. Koleva, R. Arsov, W. van der Star, M.S.M. Jetten, and M. Strous. 2006. Adaptation of a freshwater anammox population to high salinity wastewater. *J. Biotechnol.* 126:546–553.
12. Kawagoshi, Y., Y. Nakamura, H. Kawashima, K. Fujisaki, K. Furukawa, and A. Fujimoto. 2010. Enrichment of marine anammox bacteria from seawater-related samples and bacterial community study. *Water Sci. Technol.* 61:119–126.
13. Kindaichi, T., T. Ito, and S. Okabe. 2004. Ecophysiological interaction between nitrifying bacteria and heterotrophic bacteria in autotrophic nitrifying biofilms as determined by microautoradiography-fluorescence *in situ* hybridization. *Appl. Environ. Microbiol.* 70:1641–1650.
14. Kindaichi, T., I. Tsushima, Y. Ogasawara, M. Shimokawa, N. Ozaki, H. Satoh, and S. Okabe. 2007. *In situ* activity and spatial organization of anaerobic ammonium-oxidizing (anammox) bacteria in biofilms. *Appl. Environ. Microbiol.* 73:4931–4939.
15. Könneke, M., A.E. Bernhard, J.R. de la Torre, C.B. Walker, J.B. Waterbury, and D.A. Stahl. 2005. Isolation of an autotrophic ammonia-oxidizing marine archaeon. *Nature* 437:543–546.
16. Kuypers, M.M.M., A.O. Sliemers, G. Lavik, M. Schmid, B.B. Jorgensen, J.G. Kuenen, J.S.S. Damste, M. Strous, and M.S.M. Jetten. 2003. Anaerobic ammonium oxidation by anammox bacteria in the Black Sea. *Nature* 422:608–611.
17. Ludwig, W., O. Strunk, R. Westram, *et al.* 2004. ARB: A software environment for sequence data. *Nucleic Acids Res.* 32:1363–1371.
18. Lueders, T., and M. Friedrich. 2000. Archaeal population dynamics during sequential reduction processes in rice field soil. *Appl. Environ. Microbiol.* 66:2732–2742.
19. Maidak, B.L., G.J. Olsen, N. Larsen, R. Overbeek, M.J. McCaughey, and C.R. Woese. 1997. The RDP (Ribosomal Database Project). *Nucleic Acids Res.* 25:109–110.
20. Nakajima, J., M. Sakka, T. Kimura, K. Furukawa, and K. Sakka. 2008. Enrichment of anammox bacteria from marine environment for the construction of a bioremediation reactor. *Appl. Microbiol. Biotechnol.* 77:1159–1166.
21. Neef, A., R. Amann, H. Schlesner, and K.-H. Schleifer. 1998. Monitoring a widespread bacterial group: *in situ* detection of planctomycetes with 16S rRNA-targeted probes. *Microbiology* 144:3257–3266.
22. Nunoura, T., F. Inagaki, M.E. Delwiche, F.S. Colwell, and K. Takai. 2008. Subseafloor microbial communities in methane hydrate-bearing sediment at two distinct locations (ODP Leg204) in the Cascadia Margin. *Microbes Environ.* 23:317–325.
23. Okabe, S., H. Satoh, and Y. Watanabe. 1999. *In situ* analysis of nitrifying biofilms as determined by *in situ* hybridization and the use of microelectrodes. *Appl. Environ. Microbiol.* 65:3182–3191.
24. Saitou, N., and M. Nei. 1987. The neighbor-joining method: A new method for reconstructing phylogenetic trees. *Mol. Biol. Evol.* 4:406–425.
25. Schmid, M., K. Walsh, R. Webb, *et al.* 2003. *Candidatus* “Scalindua brodae”, sp. nov., *Candidatus* “Scalindua wagneri”, sp. nov., two new species of anaerobic ammonium oxidizing bacteria. *Syst. Appl. Microbiol.* 26:529–538.
26. Schmid, M.C., B. Maas, A. Dapena, *et al.* 2005. Biomarkers for *in situ* detection of anaerobic ammonium-oxidizing (anammox) bacteria. *Appl. Environ. Microbiol.* 71:1677–1684.
27. Sliemers, A.O., K.A. Third, W. Abma, J.G. Kuenen, and M.S.M. Jetten. 2003. CANON and Anammox in a gas-lift reactor. *FEMS Microbiol. Lett.* 218:339–344.
28. Stackebrandt, E., and B.M. Goebel. 1994. A place for DNA-DNA reassociation and 16S ribosomal-RNA sequence-analysis in the present species definition in bacteriology. *Int. J. Syst. Bacteriol.* 44:846–849.
29. Stahl, D.A., and R. Amann. 1991. Development and application of nucleic acid probes, p. 205–248. *In* E. Stackebrandt and M. Goodfellow (ed.), *Nucleic Acid Techniques in Bacterial Systematics*. John Wiley & Sons Ltd., Chichester, England.
30. Strous, M., J.J. Heijnen, J.G. Kuenen, and M.S.M. Jetten. 1998. The sequencing batch reactor as a powerful tool to study very slowly growing micro-organisms. *Appl. Microbiol. Biotechnol.* 50:589–596.
31. Strous, M., J.A. Fuerst, E.H.M. Kramer, S. Logemann, G. Muyzer, K.T. van de Pas-Schoonen, R. Webb, J.G. Kuenen, and M.S.M. Jetten. 1999. Missing lithotroph identified as new planctomycete. *Nature* 400:446–449.
32. Takai, K., and K. Horikoshi. 1999. Genetic diversity of archaea in deep-sea hydrothermal vent environments. *Genetics* 152:1285–1297.
33. Tsushima, I., Y. Ogasawara, T. Kindaichi, H. Satoh, and S. Okabe. 2007. Development of high-rate anaerobic ammonium-oxidizing (anammox) biofilm reactors. *Water Res.* 41:1623–1634.
34. van de Graaf, A.A., P. de Bruijn, L.A. Robertson, M.S.M. Jetten, and J.G. Kuenen. 1996. Autotrophic growth of anaerobic ammonium-oxidizing micro-organisms in a fluidized bed reactor. *Microbiology* 142:2187–2196.
35. van de Vossenberg, J., J.E. Rattray, W. Geerts, B. Kartal, L. van Niftrik, E.G. van Donselaar, J.S.S. Damsté, M. Strous, and M.S.M. Jetten. 2008. Enrichment and characterization of marine anammox bacteria associated with global nitrogen gas production. *Environ. Microbiol.* 10:3120–3129.
36. Ward, B.B., A.H. Devol, J.J. Rich, B.X. Chang, S.E. Bulow, H. Naik, A. Pratihary, and A. Jayakumar. 2009. Denitrification as the dominant nitrogen loss process in the Arabian Sea. *Nature* 461:78–81.
37. Woebken, D., P. Lam, M.M.M. Kuypers, S.W.A. Naqvi, B. Kartal, M. Strous, M.S.M. Jetten, B.M. Fuchs, and R. Amann. 2008. A microdiversity study of anammox bacteria reveals a novel *Candidatus* Scalindua phylotype in marine oxygen minimum zones. *Environ. Microbiol.* 10:3106–3119.
38. Zhong, D., E.W. Alm, D.A. Stahl, and L. Raskin. 1996. Characterization of universal small-subunit rRNA hybridization probes for quantitative molecular microbial ecology studies. *Appl. Environ. Microbiol.* 62:4504–4513.